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26 **Abstract**

27 Global phenomena, including urbanization, agricultural intensification and biotic  
28 homogenization, have led to extensive ecosystem degradation and species extinctions, and,  
29 consequently, a reduction in biodiversity. Yet, while it is now widely asserted in the research,  
30 policy and practice arenas that interacting with nature is fundamental to human health/well-  
31 being, there is a paucity of nuanced evidence characterizing how the living components of  
32 nature, biodiversity, play a role in this accepted truth. Understanding these human-  
33 biodiversity relationships is essential if the conservation agenda is to be aligned successfully  
34 with that of public health by policy-makers and practitioners. Here we show that an apparent  
35 ‘people-biodiversity paradox’ is emerging from the literature, comprising a mismatch  
36 between: (a) people’s biodiversity preferences and how these inclinations relate to personal  
37 subjective well-being; and, (b) the limited ability of individuals to accurately perceive the  
38 biodiversity surrounding them. Additionally, we present a conceptual framework for  
39 understanding the complexity underpinning human-biodiversity interactions.

40

41 *Keywords:* conservation biology, cultural ecosystem services, green space, human well-being,  
42 nature

43

44 **Introduction**

45 Despite considerable effort on the part of conservationists, the biodiversity (box 1) extinction  
46 crisis shows no sign of abating with human activities driving species losses worldwide  
47 (Cardinale et al. 2012). Solutions to stemming biodiversity loss will thus depend on changing  
48 people’s attitudes and behavior (Fuller and Irvine 2010, Duraiappah et al. 2013). Yet, the  
49 same global changes that threaten species and ecosystems, such as urbanization, agricultural  
50 intensification and biotic homogenization, also modify the ways in which humans interact

51 with nature in their day-to-day lives (Turner et al. 2004, Pilgrim et al. 2008). Human-nature  
52 interactions can be *intentional* (e.g. going to a park to feed birds, drawing trees *in-situ* within  
53 a woodland), *incidental* (e.g. running across a beach and suddenly realising you have been  
54 hearing birds calling, kicking up dead leaves as you walk although you are not cognisant of  
55 what you are doing at the time) or *indirect* (e.g. looking at images of butterflies in a book,  
56 watching a television documentary on brown bears, looking through a window to view a fox  
57 in the garden) (Keniger et al. 2013). In the highly urbanized societies which predominate in  
58 the developed, and increasingly developing, world, the human-nature interactions that occur  
59 are often restricted to green spaces (e.g. public parks and woodlands, riparian areas, private  
60 gardens; box 1) within towns and cities (Fuller and Irvine 2010). Consequently, a number of  
61 authors have argued that people are becoming progressively ‘disconnected’ from nature (e.g.  
62 Pyle 1978, Miller 2005).

63

64 The erosion of human-nature/biodiversity interactions is concerning for two reasons. Firstly,  
65 such interactions are known to provide people with multiple benefits for health/well-being  
66 (Irvine and Warber 2002, Keniger et al. 2013, Hartig et al. 2014, Lovell et al. 2014; box 1).  
67 Secondly, some authors posit that an absence of contact with nature/biodiversity could  
68 contribute towards a lack of public interest and involvement in conservation (Miller 2005).  
69 Nonetheless, the first of these points may present an important opportunity for  
70 conservationists to leverage more support for policy and management interventions to protect  
71 and enhance biodiversity, thereby improving the frequency and/or quality of people’s  
72 interactions with nature (Clark et al. 2014, Shwartz et al. 2014a). If these opportunities can be  
73 capitalized on they might bestow additional positive co-benefits by increasing public  
74 engagement in conservation.

75

76 The prevalence and costs associated with treating poor mental health and non-communicable  
77 diseases (e.g. diabetes, cardiovascular disease, depression) are expanding worldwide,  
78 particularly in developed nations (WHO 2014). As such, the beneficial outcomes associated  
79 with human-nature/biodiversity interactions (e.g. stress reduction, Peschardt and Stigsdotter,  
80 2013; improved physical exercise, Pretty et al. 2005; lower depression, Marselle et al. 2014)  
81 which can help in combatting these issues are of interest to the health sector (Coutts et al.  
82 2014). Through carefully targeted interventions, such as strategically optimizing access to  
83 urban green spaces of high ecological quality across heavily populated landscapes, relatively  
84 small gains at an individual level could scale-up to substantial cost-effective benefits across  
85 entire populations, even in comparison to approaches focused specifically on people with  
86 higher health risks (Dean et al. 2011). Investment in biodiversity could therefore be  
87 considered a worthwhile societal prophylactic, reducing the economic and human costs of ill  
88 health (Sandifer et al. 2015).

89  
90 Given that practitioners and policy-makers tasked with managing human-dominated  
91 landscapes have to deliver, and trade-off between, multiple biodiversity, individual and  
92 societal benefits (Reyers et al. 2012), environmental interventions that deliver mutually  
93 reinforcing outcomes for both biodiversity conservation and people are highly desirable.  
94 Before such scenarios can be pushed forwards, it is vital to understand the role played by  
95 biodiversity *per se*, rather than the more nebulously defined nature, in producing measurable  
96 health/well-being benefits for individuals and, in turn, the wider population. In this paper, we  
97 discuss the complex relationship between biodiversity and human health/well-being, which is  
98 emerging from a growing international literature (e.g. Lovell et al. 2014), highlighting the  
99 ‘people-biodiversity paradox’ (Fuller and Irvine 2010, Shwartz et al. 2014b pg. 87).  
100 Additionally, we present a conceptual framework that, like others in the ecological public

101 health paradigm (Coutts et al. 2014), can be a useful tool in communicating these concepts  
102 across the different research disciplines required to unpack this paradox. The people-  
103 biodiversity paradox differs conceptually from the ‘environmentalists’ paradox’ (Raudsepp-  
104 Hearne et al. 2010) in terms of both scale (the former is at the level of the individual, whereas  
105 the latter is global) and what is being measured (individual perceptions/subjective well-being  
106 in response to personal interactions with biodiversity versus objective well-being and the state  
107 of ecosystem service provision).

108

### 109 **How does biodiversity underpin human well-being?**

110 Despite ecosystem assessments being the prominent lens through which nature is valued and  
111 incorporated into decision-making (MA 2005, UKNEA 2011), our knowledge of how  
112 biodiversity underpins ecosystem functioning and services remains limited (Mace et al. 2012).  
113 This is especially true for non-material cultural ecosystem services (e.g. aesthetics, spiritual  
114 enrichment, recreation, reflection), where the relationships have rarely been investigated  
115 (Cardinale et al. 2012). How biodiversity underpins mental and physical health is less clear  
116 still and has proven harder to quantify reliably (Clark et al. 2014).

117

118 Few studies directly consider how variation in the ‘quality’ of environmental spaces, as  
119 measured by ecologists, impacts upon human well-being and individual preferences for  
120 certain elements of biodiversity (see Lovell et al. 2014 for a review). For example,  
121 epidemiological research has typically considered the size and distribution of green space  
122 surrounding properties, and the influence this has on the health/well-being of an individual  
123 (e.g. de Vries et al. 2003, Mitchell and Popham 2008). While this work provides valuable  
124 insights regarding green space accessibility/proximity across a population and the associated  
125 health/well-being benefits this might confer, it assumes that the spaces are homogenous

126 entities and does not tease apart ecological complexity in terms of, for instance, species  
127 richness (box 1), community assemblages or land cover diversity (Wheeler et al. 2015).  
128 Indeed, we know little about which aspects of biodiversity trigger the positive human well-  
129 being benefits reported in studies to-date. Furthermore, it is highly improbable that all species  
130 and ecological traits, and different compositions of these various attributes, will be  
131 advantageous or deleterious for health/well-being, particularly as responses are likely to be  
132 moderated by an array of contextual, social and cultural filters. Future research should thus  
133 explicitly consider measures of ecological quality alongside individual health/well-being  
134 outcomes.

135

136 Studies that have examined objective metrics of biodiversity (e.g. species richness and  
137 abundance) are inconclusive, identifying an inconsistent and complex relationship between  
138 biodiversity and self-reported human health/well-being. They reveal a ‘people-biodiversity  
139 paradox’ (Fuller and Irvine 2010, Shwartz et al. 2014b pg. 87), comprising a mismatch  
140 between: (a) people’s biodiversity preferences and how these inclinations relate to personal  
141 subjective well-being; and, (b) the limited ability of individuals to accurately perceive the  
142 biodiversity surrounding them.

143

144 Several papers highlight people’s preferences for greater species richness, a finding that has  
145 been repeated across a range of habitats including urban gardens (Lindemann-Matthies and  
146 Marty 2013), grasslands (Lindemann-Matthies et al. 2010a), green roofs (Fernandez-Cañero  
147 et al. 2013) and in bird song (Hedblom et al. 2014). Fuller et al. (2007) found that self-  
148 reported psychological well-being was associated positively with plant species richness, and  
149 that people could perceive accurately levels of diversity for this taxon, although this  
150 relationship was less evident for birds and not found for butterflies. Dallimer et al. (2012)

151 found no consistent relationship between plant or butterfly species richness and self-reported  
152 psychological well-being within in urban riparian environmental spaces, although a positive  
153 trend was apparent for avian diversity. Intriguingly, however, well-being was positively  
154 related to the perceived richness of all three taxonomic groups. A similar inconsistency was  
155 noted by Shwartz et al. (2014b) who discovered that people could not detect increases in  
156 flowering plant, bird or pollinator richness after experimental manipulations within public  
157 gardens, and underestimated considerably levels of diversity. Nonetheless, individuals  
158 expressed a strong preference for species richness in these green spaces and related the  
159 presence of diversity to their well-being. At a neighborhood scale, Luck et al. (2011) found a  
160 strong positive relationship between vegetation cover and self-reported well-being. However,  
161 the authors found demographic characteristics explained a greater proportion of the variation  
162 in well-being.

163

164 The people-biodiversity paradox is also evident within the literature examining individual's  
165 landscape preferences and attitudes towards biodiversity. For example, when investigating  
166 attitudes towards field margins in Swiss agricultural landscapes, Junge et al. (2009) found that  
167 people expressed a greater appreciation for margins where they estimated plant species  
168 richness was higher. Yet, actual plant richness of the field margins did not influence  
169 appreciation. Thus, as was true of the urban green space studies highlighted above, people's  
170 predilections appear to be driven by the biodiversity they perceive to be present. However,  
171 there are exceptions. Qiu et al. (2013) discovered that people could correctly estimate the  
172 differences in plant diversity across habitats, and that the species richness of this taxon was  
173 not related to preference, with open park locations rated more highly than areas of more  
174 complex vegetation. Likewise, Shanahan et al. (2015a) found that people do not preferentially



175 visit parks with higher tree and vegetation cover, despite these areas having the potential for  
176 enhanced experiences of biodiversity.

177

178 The disparities outlined above may be a consequence of ecological factors such as spatial  
179 scale, taxonomic group and the metrics used to measure biodiversity. Findings at a broad  
180 scale (i.e. asking people to rank images of landscapes by the level of human disturbance)  
181 indicate that people can reliably identify differences in landscape intactness (Bayne et al.  
182 2012), but fail to estimate the objective level of greenness of their neighborhood (Leslie et al.  
183 2010). While Lindemann-Matthies et al. (2010b) reported a positive relationship between  
184 plant species richness and individual aesthetic preferences, the effect was modified by the  
185 spatial distribution of the plants. Additionally, plant communities consisting of the same  
186 number of species were perceived to be more species-rich when evenness (the relative  
187 abundance of different species) was higher (Lindemann-Matthies et al. 2010b). This suggests  
188 that species richness alone may not be the best measure of biodiversity when considering  
189 human responses to, and appreciation of, biodiversity. Indeed, this is understandable, as many  
190 species cannot be detected without specialist training (e.g. because they are difficult to  
191 identify) or without a great deal of effort (e.g. because of their elusive behavior). When  
192 unpicking the people-biodiversity paradox, researchers should consider using a suite of more  
193 resolved biodiversity metrics (e.g. abundance, evenness, functional diversity) to determine the  
194 ecological quality of environmental green spaces (Lovell et al. 2014).

195

### 196 **Explicit consideration of the complexity associated with human well-being and** 197 **biodiversity**

198 It is possible that the emerging people-biodiversity paradox is a result of the multi-  
199 dimensionality of both biodiversity and human well-being, making it difficult to account for

200 and measure the complex social and ecological characteristics that may influence the outcome  
201 of interactions (Hartig et al. 2014, Lovell et al. 2014). The concepts of health and well-being  
202 are just as multifarious as that of ecological quality, incorporating a wealth of different  
203 aspects of human physiological, cognitive, emotional, social and spiritual wellness, and  
204 studies have explored these facets from several disciplinary perspectives (Irvine and Warber  
205 2002, Keniger et al. 2013, Irvine et al. 2013). Heterogeneity in research design, and the use of  
206 different ecological and well-being measures, thus reflects the complexity that social and  
207 natural scientists are grappling with in trying to understand how people derive benefits from  
208 interacting with nature/biodiversity. Our conceptual framework (figure 1) illustrates that such  
209 interactions could generate outcomes for an individual's health/well-being and, in turn, this  
210 might relate to human perceptions of, and behaviors towards, biodiversity.

211

212 The type and intent of the human-biodiversity interaction are likely to influence the outcome  
213 (Church et al. 2014), which might be positive, neutral or negative (figure 1). Additionally,  
214 experiences of biodiversity can be influenced by physical/environmental characteristics  
215 associated with the point of interaction, such as the season and prevailing weather conditions  
216 (figure 1, table 1). These filters are often ignored in research projects, but are potentially  
217 important determinants of outcomes (White et al. 2014). While the majority of studies  
218 conducted on human-nature/biodiversity interactions thus far have concentrated on benefits  
219 gained by people, disservices also require research attention (Dunn 2010), as practitioners and  
220 policy-makers need to be able to make fully informed decisions in a land-use planning and  
221 management context (Lyytimäki and Sipilä 2009). At the most extreme, interactions with  
222 biodiversity can lead to death and injury, for instance, through attacks from predators or via  
223 the contraction of pathogens. Human-wildlife conflict can also lead to diminished health/well-  
224 being in addition to physical injury or pathology (Barua et al. 2013) and, in an urban context,

225 close contact with nature has been associated with fear, disgust and discomfort (Bixler and  
226 Floyd 1997).

227

228 The outcome of an interaction with biodiversity can feedback to the individual (figure 1),  
229 changing aspects of their ecological knowledge, values, and underlying health/well-being.  
230 Indeed, a particular interaction might be perceived as positive or negative, depending on the  
231 individual making the evaluation (Buchel and Frantzeskaki 2015). In turn, this could  
232 contribute to the likelihood that the individual will subsequently interact with biodiversity and  
233 may influence future outcomes (e.g. positive interactions might predispose future outcomes to  
234 being more positive and vice versa). A suite of individual characteristics can moderate both  
235 the magnitude and direction of an outcome, as well as the probability that an interaction will  
236 take place (figure 1, table 2). To illustrate, a review of fear of crime experienced in urban  
237 green spaces found variability in responses according to factors such as age, gender, socio-  
238 economic status, frequency of visits and familiarity with the site, as well as the bio-physical  
239 attributes of the areas (Maruthaveeran and van den Bosch 2014). Cultural factors are also  
240 likely to be important. A recent paper by Lindemann-Matthies et al. (2014) demonstrated that  
241 a cohort of Chinese people did not show a preference for biodiverse forest, whereas the  
242 comparative Swiss participants favored species rich forest over monoculture. Similarly, a  
243 study in Singapore found that neither access to, nor use of, green spaces influenced measures  
244 of well-being (Saw et al. 2015). There is a paucity of such cross-cultural studies, with most  
245 work on human-nature/biodiversity interactions being geographically biased towards  
246 industrialized regions of the Global North (Keniger et al. 2013). This hinders our  
247 understanding, and there is a need for greater focus on biodiversity rich countries where urban  
248 development is accelerating rapidly (Lindemann-Matthies et al. 2014).

249

250 How frequently people choose to visit green spaces, if at all, can be influenced by both the  
251 characteristics of individuals (table 2), as well as the accessibility/proximity of the green  
252 space (table 1). The contribution of these different sets of attributes appears to be variable,  
253 with contradictory results reported in studies. For example, people's nature orientation, that  
254 is, the affective, cognitive and experiential relationship they have with the natural world, has  
255 been shown by some to be more important in determining time spent in urban green spaces  
256 than the availability of nearby green space (Lin et al. 2014). Conversely, others report that  
257 proximity and the time it takes individuals to reach a site are stronger predictors of visit  
258 frequency (Dallimer et al. 2014). The visit duration can also influence the outcome of  
259 interactions (a dose-response relationship), with research typically finding a positive  
260 relationship between the time spent in a green space and the response (White et al. 2013).  
261 However, others have found less straightforward dose-response relationships. For instance,  
262 Barton and Pretty (2010) found diminishing, but still positive, mental health returns from  
263 higher intensity and duration green exercise, while Shanahan et al. (2015b) suggests several  
264 potential dose-response relationships.

265

266 A further complexity that requires careful consideration is that spending time in green spaces  
267 can be beneficial to individuals, not necessarily because of interaction with biodiversity, but  
268 by virtue of the fact it encourages and facilitates behaviors that are known to be mentally and  
269 physical favorable, such as exercise and social interaction. It is therefore important to evaluate  
270 the extent to which human-biodiversity interactions provide added value. Research into green  
271 exercise, for example, has shown that there are synergistic benefits associated with taking part  
272 in physical activities while viewing nature (Pretty et al. 2005).

273

274 **What are the consequences of the people-biodiversity paradox for conservation?**

275 If, as recent studies suggest, human-biodiversity interaction outcomes are influenced by  
276 people's perceptions of biodiversity, rather than objective measures, the role of ecological  
277 knowledge in influencing the relationship is a key dimension worthy of consideration. The  
278 lack of ecological knowledge in developed world citizens (Pilgrim et al. 2008, Dallimer et al.  
279 2012) might support authors' assertions that there is a growing 'disconnection' between  
280 people and nature (Pyle 1978, Turner et al. 2004, Miller 2005). They propose that an  
281 'extinction of experience' is occurring because individuals are isolated increasingly from  
282 nature in their everyday lives and, as such, they have less impetus to protect and experience  
283 nature, leading to a vicious deleterious cycle. Social or education interventions have been  
284 advocated as a means to reverse this negative feedback. For instance, research has shown that  
285 people with more taxonomic knowledge express preferences for more species rich flower  
286 meadows (Lindemann-Matthies and Bose 2007), and children who participated in an  
287 educational program had an increased appreciation of local nature (Lindemann-Matthies  
288 2005). However, questions remain as to whether such interventions have a long-term impact  
289 on levels of interest and engagement with biodiversity (Shwartz et al. 2012).

290

291 If people are only responding positively to certain traits and assemblages of species, it is  
292 possible that these might not be the biodiversity elements that conservationists would wish to  
293 support. Urban areas are highly susceptible to biotic homogenization and harbor many non-  
294 native species (McKinney 2002). As yet, it is still unclear whether the nativeness of species  
295 makes a difference to the well-being response an individual receives from an interaction.  
296 People may value species that they know to be native more (Lundhede et al. 2014), although  
297 non-native species may possess traits (e.g. larger body size, more colorful or behaviorally  
298 distinct) which people prefer (Frynta et al. 2010). This could present a potential challenge and  
299 conflict for conservationists and practitioners, who may seek to promote native taxa through

300 the management of non-native species, but also need to encourage the health/well-being  
301 benefits that may gained from interacting with charismatic non-native species. A better  
302 understanding of the public perception of non-native species could feed usefully into the on-  
303 going debates on the legitimacy of the novel ecosystem (box 1) concept (Hobbs et al. 2006,  
304 Kowarik 2011), as well as providing an evidence-base for land-use planning, management  
305 and decision-making.

306

307 Even if future research continues to corroborate the advantages people can gain from  
308 interacting with biodiversity, individuals might not consciously relate these benefits to  
309 biodiversity *per se*. If this is the case, there is no reason to expect an individual's perception  
310 of biodiversity to alter as a consequence human-biodiversity interactions and, subsequently, to  
311 presume a shift towards more pro-biodiversity behavior. Indeed, positive attitudes towards  
312 biodiversity alone do not translate into pro-biodiversity behaviors (Waylen et al. 2009) (figure  
313 1), being modified by numerous external as well as internal factors, including subjective  
314 norms, facilitating factors and moral obligations (Clayton and Myers 2009). Much more  
315 research is needed to discern the links between exposure to biodiversity and how this might,  
316 ultimately, lead to shifts in underlying attitudes and behavior. Beyond education,  
317 understanding what individual's perceive as constituting a preferable biodiverse environment  
318 will allow for human-modified landscapes to be designed in a manner which delivers benefits  
319 to both people and biodiversity.

320

## 321 **Conclusion**

322 The examples presented here of the people-biodiversity paradox illustrate the need for careful  
323 consideration before a straightforward relationship between increased biodiversity and  
324 improved human well-being can be implied. If we wish to align the agendas of public health

325 and biodiversity conservation, we first need to understand the mechanisms behind the people-  
326 biodiversity paradox, and the added value that enhanced people-biodiversity interactions can  
327 deliver for conservation. Well-designed and carefully conducted interdisciplinary research,  
328 which genuinely bridges traditional disciplinary boundaries, will be the key to effectively  
329 unpacking this paradox.

330

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337

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**Figure 1:** Conceptual framework of human-biodiversity interactions and potential outcomes for health and well-being, perceptions of biodiversity and pro-biodiversity behavior. Human-biodiversity interactions can lead to a cascade of potential outcomes. The question marks represent less well-understood relationships. The dotted lines represent feedbacks from outcomes back to biodiversity or the individual.



## Box 1: Key terminology

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Biodiversity	The variability among living organisms from all sources including, <i>inter alia</i> , terrestrial, marine and other aquatic ecosystems and the ecological complexes of which they are part; this includes diversity within species, between species and of ecosystems	Convention on Biological Diversity ( <a href="http://www.cbd.int/convention/articles/default.shtml?a=cbd-02">www.cbd.int/convention/articles/default.shtml?a=cbd-02</a> )
Green space	Open, undeveloped land with natural vegetation	Centers for Disease Control and Prevention ( <a href="http://www.cdc.gov/healthyplaces/terminology.htm">www.cdc.gov/healthyplaces/terminology.htm</a> )
Novel ecosystem	Ecosystems which have been heavily modified by humans, and differ in composition and/or function from present and past systems	Hobbs et al. 2009
Human health	Health is ‘a complete state of physical, mental and social well-being, and not merely the absence of disease or infirmity	World Health Organization WHO 1948
Human well-being	(Subjective) well-being encompasses different aspects – cognitive evaluations of one’s life, happiness, satisfaction, positive emotions such as joy and pride and negative emotions such as pain and worry	Stiglitz et al. 2009
Species richness	The number of species observed in a defined geographic location	Begon et al. 2006

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**Table 1:** Illustrative physical/environmental characteristics which could influence the likelihood that people will interact with nature/biodiversity, and the outcome of such interactions.

Characteristic	Description and supporting examples
Season	Seasonal changes affect the well-being of office workers (Hitchings 2010)
Weather	Landscape preferences are influenced by climatic conditions (White et al. 2014)
Accessibility	People who report that they have easy access to green spaces use green spaces more regularly (Hillsdon et al. 2011)
Proximity	People with less green space in close proximity to their home reported greater loneliness and a perceived shortage of social support (Maas et al. 2009). Populations exposed to the greenest environments have the lowest levels of health inequalities (Mitchell and Popham 2008). People visit more frequently when it takes less time to reach a green space (Dallimer et al. 2014)

**Table 2:** Illustrative individual characteristics which could influence the likelihood that people will interact with nature/biodiversity, and the outcome of such interactions.

Characteristic	Description and supporting examples
Gender	Gender differences have been observed in associations between urban green space and health outcomes (Richardson and Mitchell 2010). Women demonstrate a preference for higher plant species richness than men (Lindemann-Matthies and Bose 2007, Lindemann-Matthies et al. 2010a)
Age	Proximity to green space has a greater influence on the health of the elderly than other age groups (de Vries et al. 2003). Older people prefer species rich field margins (Junge et al. 2009) and meadows (Lindemann-Matthies and Bose 2007)
Education	Health benefits from proximity to green space are greater for people with a lower level of completed formal education (de Vries et al. 2003)
Socio-demographic/ economic factors	There are racial and economic inequalities regarding access to biodiversity, for example fewer native birds have been found in neighborhoods comprising of predominantly Hispanic and lower-income people (Lerman and Warren 2012)
Home location	People who identify themselves as ‘urban’ report lower levels of restoration from images of nature than ‘rural’ individuals (Wilkie and Stavridou 2013)
Culture	Chinese study participants demonstrate no strong preferences for biodiversity when compared to Swiss participants, who favored species-rich forests over monocultures (Lindemann-Matthies et al. 2014). The wellbeing of residents in Singapore was not affected by access to, or the use of, green spaces (Saw et al. 2015)
Childhood experience	People who spent their childhood in a more natural environment show a greater preference for green roofs over gravel (Fernandez-Cañero et al. 2013)
Connectedness to nature	Residents living in neighborhoods with greater richness and abundance of bird species and density of plants had a higher connection to nature (Luck et al. 2011)
Ecological knowledge	Children who participated in an educational program had increased appreciation of local nature (Lindemann-Matthies 2005). People with better wildlife identification skills were able to more accurately estimate the species richness of surrounding vegetation, birds and butterflies (Dallimer et al. 2012)
Intention	Although interacting with nature is beneficial to urban park visitors, it was not a main motivation for visiting (Irvine et al. 2013). Frequent users of urban green spaces state motivations relating to physical activities, whereas infrequent users motivations are more associated to the quality of the space (Dallimer et al. 2014)
Social interaction	Individuals who visited natural areas accompanied by children experienced less restoration than those who were alone (White et al. 2013). Fear of crime influences some individuals to avoid urban green spaces (Maruthaveeran and van den Bosch 2014)

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State of mind

Urban green spaces which are perceived to contain more nature are also perceived to be more restorative by stressed individuals (Peschardt and Stigsdotter 2013)

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