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REVIEW

The ecological outcomes of biodiversity offsets under “no net loss” policies: A global review

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Abstract

No net loss (NNL) biodiversity policies mandating the application of a mitigation hierarchy (avoid, minimize, remediate, offset) to the ecological impacts of built infrastructure are proliferating globally. However, little is known about their effectiveness at achieving NNL outcomes. We reviewed the English-language peer-reviewed literature (capturing 15,715 articles), and identified 32 reports that observed ecological outcomes from NNL policies, including >300,000 ha of biodiversity offsets. Approximately one-third of NNL policies and individual biodiversity offsets reported achieving NNL, primarily in wetlands, although most studies used widely criticized area-based outcome measures. The most commonly cited reason for success was applying high offset multipliers (large offset area relative to the impacted area). We identified large gaps between the global implementation of offsets and the evidence for their effectiveness: despite two-thirds of the world’s biodiversity offsets being applied in forested ecosystems, we found none of four studies demonstrated successful NNL outcomes for forested habitats or species. We also found no evidence for NNL achievement using avoided loss offsets (impacts offset by protecting existing habitat elsewhere). Additionally, we summarized regional variability in compliance rates with NNL policies. As global infrastructural expansion accelerates, we must urgently improve the evidence-base around efforts to mitigate development impacts on biodiversity.

KEYWORDS

biodiversity offsets, compliance, conservation outcomes, mitigation hierarchy, multipliers, no net loss, policy effectiveness

1 | INTRODUCTION

We are living in an age of both severe biodiversity declines and unprecedented global expansion of built infrastructure (IPBES 2019; Laurance et al., 2015). Approximately a quar-

ter of all species from red-list assessed groups are threatened with extinction, with the unmitigated impacts of infrastructure a major driver (IPBES 2019; Maxwell, Fuller, Brooks, & Watson, 2016). These impacts are expected to intensify over the coming decades, with dramatic increases in our

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transport networks, urban footprint, and energy production facilities already under way (Steffen, Broadgate, Deutsch, Gaffney, & Ludwig, 2015). Mitigating these impacts is therefore an urgent global priority. Currently, one of the most widely used tools for addressing the environmental impacts of infrastructure are No Net Loss (NNL) policies (Bennett et al., 2017), which mandate that a mitigation hierarchy (MH) is applied to sequentially avoid, minimize, remediate, and offset the biodiversity impacts of new developments (Bennett et al., 2017), with some variation among policies (e.g., U.S. mitigation sequence: avoid; minimize; compensate).

NNL policies are proliferating around the world (Bennett et al., 2017), reflected in the widespread implementation of biodiversity offsets (Bull & Strange, 2018). Throughout, we use the term “biodiversity offsets” to refer to all offsets implemented as the final stage of NNL policies, as nearly all policies focus on achieving outcomes that are related to or underpinned by biodiversity. However, the exact ecological characteristics for which these policies aim to achieve NNL vary considerably (e.g., U.S. wetland compensatory mitigation protects “wetland acreage and function” (EPA, 2008)). There is a notable lack of evidence regarding the actual outcomes of NNL policies because of the relative immaturity of many policies, a lack of data transparency surrounding NNL implementation (Bull et al., 2018), and challenges evaluating largely unobservable outcomes of the MH process (e.g., identifying avoided impacts; Sinclair, 2018). Much of the evidence of NNL effectiveness comes from individual offset case studies or simulation studies (e.g., Sonter, Tomsett, Wu, & Maron, 2017; Thorn, Hobbs, & Valentine, 2018). In the absence of a coherent body of evidence regarding actual outcomes, many theoretical criticisms and defenses of NNL have been discussed in the literature. Criticisms revolve around the ecological feasibility of restoration (Maron et al., 2012), choice, and definition of biodiversity “units” (Bull, Suttle, Gordon, Singh, & Milner-Gulland, 2013), perverse incentives to game offset policies through manipulation of counterfactuals (Gordon, Bull, Wilcox, & Maron, 2015), ethics of biodiversity trading (Ives & Bekessy, 2015), and the weakening of institutions that safeguard the environment (Walker, Brower, Stephens, & Lee, 2009). In response, defenses of NNL acknowledge that well-targeted infrastructural expansion can deliver considerable well-being benefits, and when applied according to best practice (Bennett et al., 2017; Bull et al., 2013), NNL can facilitate this without damaging biodiversity overall. Furthermore, NNL buffers impacts on biodiversity that would most likely occur anyway in the absence of NNL policy (von Hase & ten Kate, 2017). Additionally, the organization and financing of offsets may make avoiding impacts initially more favorable to developers (Calvet, Napoléone, & Salles, 2015). However, without an empirically grounded

evidence base, it is unclear which arguments dominate in practice.

Evidence from case studies shows that NNL policies result in both successes and failures (Quigley & Harper, 2006a). As with any conservation intervention, developing evidence about the contextual factors that predict NNL success is essential. Additionally, researchers have reviewed and tested the major indicators of biodiversity proposed for use in NNL and evaluated whether they provide useful approximations of biodiversity changes (Bezombes, Gaucherand, Spiegelberger, Gouraud, & Kerbiriou, 2018). However, little work has synthesized which indicators are used in the practical implementation of NNL globally.

Several high-profile NNL policies have now been implemented for sufficient timescales for a preliminary understanding of outcomes to emerge (e.g., Gibbons, Macintosh, Constable, & Hayashi, 2018). However, there remains no synthesis of all the information available on the actual observed outcomes of NNL policies from around the world (i.e., whether they have demonstrably achieved NNL of their ecological characteristic of interest). Addressing this, we reviewed the global literature on the outcomes of NNL policies to synthesize literature gaps and coverage, summarize the state of the knowledge on the determinants NNL outcomes, assess the biodiversity metrics used in practice, assess regional compliance with NNL policies, and evaluate the validity of the existing literature. For clarity, our study addresses both the effectiveness of NNL policies (i.e., the application of the MH to development impacts under jurisdiction of a NNL policy) and individual biodiversity offsets (i.e., whether or not offsets achieve NNL in chosen biodiversity indicators at project scales).

2 | METHODS

2.1 | Review protocol

We conducted a rapid evidence assessment (Khangura, Konnyu, Cushman, Grimshaw, & Moher, 2012) of peer-reviewed literature on NNL outcomes. Our search term (Supporting Information) comprised a set of strings linked by Boolean operators describing the following:

- alternative offset types (e.g., “environmental”),
- “offset” and commonly used alternatives (e.g., “compensat*”),
- impact evaluation (e.g., “outcome*”),
- and excluding nuisance terms (refined by identifying unrelated papers in the first 200 hits of our Web of Science review; e.g., “gas mitigation”)

Performing the same search in Web of Science and Scopus databases (final search date March 13, 2019), we removed

repeats and then reviewed the remaining studies using the “metagear” package in R (Lajeunesse, 2016; R Core Team 2018). We conducted a first assessment of potentially relevant literature by selecting all studies mentioning NNL policies or offsets in their abstracts, then read the full papers to identify whether our inclusion criteria were met. We limited our review to studies published from 2003 to 2019, to account for the major reforms to the effectiveness of U.S. wetland mitigation policy introduced by the National Wetlands Mitigation Action Plan in December 2002 (Hough & Robertson, 2009). We restricted our search to English-language articles from relevant topic categories (Supporting Information). Previous research has shown that English captures most literature on offsets tied to international funding requirements, studies from North America and Oceania, and a substantial proportion of European literature, so our findings should be representative of the global literature (Bull & Strange, 2018; Bull et al., 2018). Additionally, we searched through all reference lists in papers meeting our inclusion criteria for additional literature.

2.2 | Data extraction

Papers were included in our database if they reported observed (i.e., not simulated) ex-post ecological or land cover-related outcomes of policies with an explicit NNL-or-better objective for aspects of biodiversity. We limited our search to peer-reviewed publications only (including conference proceedings and book chapters) to attempt to overcome the data quality issues highlighted by other reviews of offset studies that include the gray literature (Theis et al., 2019), but recognize that the majority of NNL implementation occurs outside academic evaluation. Papers reporting evaluations of individual offset projects were included if they specified the impacts (as a minimum defining the impacted habitat and area) associated with the offsets, thus allowing for a rudimentary assessment of biodiversity losses and gains. These papers compared biodiversity at offsets with either biodiversity at the impacted site (pre-initiation of impacts), or with a biodiversity reference site (Table S2 for studies considered but ultimately rejected). Notably, while we included only these studies that allowed for a site-specific estimate of biodiversity losses and gains and thus a basic evaluation of whether NNL was achieved, some key NNL policies do not assess biodiversity losses and gains in this way (e.g., U.S. wetland mitigation policy mandates that compensation sites achieve benchmark ecological criteria rather than explicitly achieving the same level of ecosystem functioning as impacted wetlands). Therefore, such NNL policies may in theory achieve full compliance but not NNL.

For each study/individual offset project where possible we extracted information regarding the following:

- type of biodiversity outcome variable used to assess losses and gains,
- magnitude of the outcome variable at the offset and impact/control site,
- affected type of biodiversity (e.g., forest, species),
- location,
- mean offset age (mean time between offset initiation and outcome evaluation),
- spatial scale (Table 1),
- whether NNL was achieved for the outcome variable of interest,
- and article author’s explanations for why/why not (including only reasons that addressed the specific outcome variable used).

For each reported outcome variable, we assigned it the appropriate level for four descriptive categories (Table 1). If a paper reported multiple ecological indices or outcome variables, we recorded them all. For individual offsets that presented time-series outcomes, we recorded the outcome variable at the latest time-period to allow the maximum time for ecological recovery in the offset-control comparison. For NNL policies presenting time-series outcomes, we took the sum of the outcome variables across time periods to capture the policy’s impact across the entire evaluation period (Table S3). We recorded information about the policy outcomes across its entire geographical jurisdiction (i.e., if a paper reported localized habitat losses but NNL overall (e.g., across an entire state), we recorded that NNL was achieved). We extracted data from figures using WebPlotDigitizer (Rohatgi, 2015). We recorded the raw values for outcome variables and used them to infer NNL outcomes, except for papers that compared outcomes between offset and impact/reference sites using statistical tests, where we used the test’s outcome to inform NNL designation. When studies reported that outcomes for some of the projects they evaluated was unknown, we recalculated the percentage of projects reporting successes and failures restricting the total sample to only projects for which the outcome was known (Table S3). For offset project studies that reported per-unit-area values for a given outcome variable, we multiplied the outcome variable for the offset site by the offset ratio so that the final comparison between biodiversity at the impact and offset sites accounted for differences in area between the two (Table S3). Therefore, for project-scale evaluations, we did not include area as an outcome variable, but for program and landscape-scale evaluations, habitat area was included as an outcome. Additionally, we noted two important aspects of offset design: whether the described offsets referenced the additionality of their associated conservation actions (i.e., whether the biodiversity gains at the offset were additional to what would have been present

TABLE 1 Categorization of information from each study evaluating outcomes from biodiversity offsets or NNL policies

Category	Groupings	Inclusion criteria
Scale	Landscape	Assess changes in the total area of a particular land cover type regulated under a regional NNL policy (although note that some individual impacts within the geographical jurisdiction of the policy will not be regulated by the policy because of legal exemptions or illegal impacts).
	Program	Assess the outcomes of a defined portfolio of offsets without necessarily comparing them with their associated impacts.
	Project	Report the results of individual impact/reference and offset pairs.
Offset type	Creation	Result in the creation of new habitat where none existed previously.
	Restoration	Restoration or enhancement of degraded habitats; may or may not result in additional habitat area.
	Protection	Protection of existing habitats, may or may not involve conservation management. No additional area for conservation.
Data type	Ecological site-based	Primary data collected on site.
	Expert judgment	Judgment about outcomes elicited from experts.
	Official documentation	Data retrieved from official documentations such as mitigation permit files or offset registries.
	Remote sensing	Use remote sensing to assess changes in habitat extent.
Outcome variable type (Table S3)	Community indices	General indices used to describe ecological communities (e.g., species richness; Simpson index). Do not account for species identity.
	Community densities	Indices showing the abundance of an aspect of biodiversity per unit area (e.g., g/m ² fish biomass).
	Habitat area	Area of habitat.
	Habitat quality	Quality of habitat (e.g., percentage coverage of vegetation types associated with the offset habitat type).
	Indices of biotic integrity	Indices of biotic integrity (Karr 1981), partially account for changes in species identity.
	Regulatory compliance	Degree to which a given compliance criterion has been met (compliance does not necessarily demonstrate the achievement of NNL).
	Species population proxy	Direct monitoring or species proxy monitoring methods targeting a particular individual or set of species (e.g., population abundance; environmental indicators of species activity levels).

in the absence of the offset), and whether losses/gains were evaluated against a static or dynamic counterfactual (Bull, Gordon, Law, Suttle, & Milner-Gulland, 2014; McKenney & Kiesecker, 2010).

We also assessed the internal validity of site-based assessments of individual biodiversity offsets, paying particular attention to potential selection bias and performance bias (Bilotta, Milner, & Boyd, 2014). We recorded information about the:

- study design (e.g., before-after-control-impact);
- control used (e.g., either impact-site or reference-site);
- sampling methods and whether those descriptions were sufficiently randomized or open to selection bias;
- and number of time periods sampled and whether this was sufficient to capture intertemporal ecological dynamics.

3 | RESULTS

3.1 | Overview of studies

Our searches returned 15,715 articles once duplicates were removed. After screening abstracts for relevance, we fully assessed 418 articles for inclusion (Table S1). Twenty-nine studies met our inclusion criteria (7% of potentially relevant studies), with a further three identified via in-article citations, leaving 32 studies from five countries (Table 1; Figure 1). Our database includes four landscape-scale, 18 program-scale, and 10 project-scale studies (covering 26 projects), and accounts for a minimum of 300,000 ha of offsets and 180,000 ha of impacts, representing approximately 2% of the global area of spatially explicit known offset implementation (Bull & Strange, 2018). In total, we identified 121 outcome variables (column 11, Table 2) from 48 NNL policies or individual offsets (1–44 outcomes per study, mean = 3.75). NNL assessments could not be made for eight studies, as the sole

TABLE 2 Outline of all studies included in our review

Study	Location	NNL policy	Mean offset age	Affected habitat/species	Dominant offset type	Total offset area (ha)	Total impact area (ha)	Data type	Outcome variables	NNL achieved?
Gibbons et al., 2018	Australia, New South Wales	New South Wales native vegetation act	10	Forested (native vegetation)	Protection	83459	21928	Remote sensing	Native vegetation cover	No
Lindenmayer et al., 2017	Australia, New South Wales	Hume Highway environmental assessment	1.5	Forest species	Restoration	NA	NA	Ecological site-based	Superb parrot and squirrel glider % nest box/tree hollow occupancy	No
Levrel et al., 2017	USA, Florida	US Clean Water Act section 404	10	Landscape Wetland	NA	58575	114575	Remote sensing	Wetland area	No
Van den Bosch & Matthews, 2017	USA, Illinois	US Clean Water Act section 404	14	Wetland	Restoration	NA	NA	Ecological site-based	Percentage of sites meeting various compliance performance standards	NA
May et al., 2017	Australia, Western Australia	Western Australia state Environmental Protection Act	11	Forested	Protection	2841	NA	Official documentation	Percentage of implemented, evaluated offsets that successfully purchase land for conservation, or achieve their completion criteria	NA
Drielsma et al., 2016	Australia, New South Wales, Lower Murray Darling catchment	New South Wales native vegetation act	7	Landscape Forested	Protection	107994	40458	Remote sensing	Natural vegetation area	No
Goldberg & Reiss, 2016	USA, Florida, Lower St John's River Basin	US Clean Water Act section 404	7	Wetland	Protection	11123	1412	Official documentation	Wetland/upland area	No
Fickas et al., 2016	USA, Oregon, Willamette Valley	US Clean Water Act section 404	12	Landscape Wetland	NA	NA	NA	Remote sensing	Wetland area	Yes
Murata & Feest, 2015	UK, Wales, Cardiff Bay	Cardiff Bay Barrage Environmental Statement	7.5	Wetland	Restoration	273	207	Ecological site-based	Various ecological indices including species richness and Simpson index	Yes
Hobbs & MacAller, 2014	USA, California, Santa Maria River	US Clean Water Act section 404	2	Wetland	Restoration	52	52	Ecological site-based	Meeting performance standards for native species richness and cover, and non-native cover	NA
Stone et al., 2013	UK, England	EU Habitats Directive; English Nature Bat mitigation guidance	NA	Urban species	Restoration	NA	NA	Official documentation	Percentage change in bat presence/absence and abundances post-development	No
Hill et al., 2013	USA, North Carolina	US Clean Water Act section 404	7	Wetland	Restoration	8000	NA	Ecological site-based	Percentage of wetland and stream components achieving regulatory compliance	NA
Pickett et al., 2013	Australia, New South Wales, Sydney	Sydney Olympic Park development licence	10	Wetland species	Creation	6.4757	0.3351	Ecological site-based	Population size	Yes
Kozich & Halvorsen, 2012	USA, Michigan, Upper Peninsula	US Clean Water Act section 404	3.5	Wetland	Creation	75.1	28.8	Ecological site-based, official documentation	Percentage of wetlands meeting invasive species performance criteria, wetland area	Yes (area), NA (compliance criteria)
Carle, 2011	USA, North Carolina	US Clean Water Act section 404	7	Landscape Wetland	NA	NA	NA	Remote sensing	Wetland area	No
Reiss et al., 2009	USA, Florida	US Clean Water Act section 404	6	Wetland	Restoration	24014	NA	Expert judgment	Percentage of banks achieved or on course to achieve final regulatory performance criteria	NA

(Continues)

TABLE 2 (Continued)

Study	Location	NNL policy	Mean offset age	Scale	Affected habitat/species	Dominant offset type	Total offset area (ha)	Total impact area (ha)	Data type	Outcome variables	NNL achieved?
Robertson & Hayden, 2015	USA, Illinois, Chicago district	US Clean Water Act section 404	NA	Program	Wetland	Restoration	1209.12	448	Official documentation	Wetland area	Yes
Kettlewell et al., 2008	USA, Ohio, Cuyahoga River watershed	US Clean Water Act section 404	NA	Program	Wetland	Creation	NA	14.95	Ecological site-based	Wetland area	Yes
Breaux et al., 2005	USA, California, San Francisco district	US Clean Water Act section 404	11.5	Program	Wetland	Restoration	135.2	40	Ecological site-based	Wetland area, percentage of projects meeting all regulatory performance criteria	Yes (area), NA (compliance criteria)
Teels et al., 2004	USA, Virginia, Warrenton	US Clean Water Act section 404	10	Project	Wetland	Creation	7.3	6	Ecological site-based	Species richness, index of biotic integrity	No
Thorn et al., 2018	Australia, Western Australia, Beeljar regional park	Western Australia state Environmental Protection Act	NA	Project	Forest species	Protection	523	97.85	Ecological site-based	Area of high quality Carnaby's cockatoo and red-tailed black Quenda diggings	No
Harper & Quigley, 2005	Canada	Fisheries Act, Habitat Policy	3.6	Program	Wetland	Creation	102.0388	41.9562	Official documentation	Fish habitat area	Yes
Quigley & Harper, 2006b	Canada	Fisheries Act, Habitat Policy	4.4	Program	Wetland	Creation	NA	NA	Ecological site-based	Percentage compliance with biological performance criteria	NA
Quigley & Harper, 2006a	Canada, Manitoba; British Columbia; Nova Scotia; Ontario; New Brunswick	Fisheries Act, Habitat Policy	5	16 Projects	Wetland	Creation	1.6781	1.9606	Ecological site-based	Periphyton biomass, invertebrate abundance, coverage of riparian vegetation, fish biomass	Mixed
Morgan & Roberts, 2003	USA, Tennessee, NA	US Clean Water Act section 404	NA	Program	Wetland	Protection	77.7	38	Ecological site-based	Wetland area, percentage of sites meeting all regulatory performance criteria	No (area), NA (compliance criteria)
Hegberg et al., 2010	USA, Indiana, Indianapolis	US Clean Water Act section 404	3	Two projects	Wetland	Creation	11575 feet	NA	Ecological site-based	Various ecological indices including index of biotic integrity	No
BenDor, Sholtes, & Doyle, 2009	USA, North Carolina	US Clean Water Act section 404	NA	Program	Wetland	Restoration	NA	NA	Official documentation	Wetland area, stream length	Yes
Shafer & Roberts, 2008	USA, Florida	US Clean Water Act section 404	21	Program	Wetland	Creation	NA	NA	Ecological site-based	Percentage of sites likely to meet regulatory permit criteria	NA
Shea et al. 2007	USA, Florida, Hillsborough County	US Clean Water Act section 404	1	Project	Wetland	Creation	400	190	Ecological site-based	Percentage of samples meeting permit criteria for wetland plant coverage and nuisance species	NA
Garland et al., 2017	UK, Somerset, Bath	EU Habitats Directive; English Nature Bat mitigation guidance	2	Project	Urban species	Creation	NA	NA	Ecological site-based	Presence of brown long-eared bat and common pipstrelle roosts	Yes (long-eared bat), No (common pipstrelle)
BenDor et al., 2007	USA, Illinois, Chicago	US Clean Water Act section 404	NA	Program	Wetland	Restoration	1053.6	617.6	Official documentation	Wetland area	Yes
Bezombes et al., 2019	France, Auvergne-Rhône-Alpes, Isère	French law on biodiversity no. 2016-1087	4	Program	Wetland, wetland species	Protection	182.76	59.79	Ecological site-based	Percentage offsets where compensation habitat or species present	No

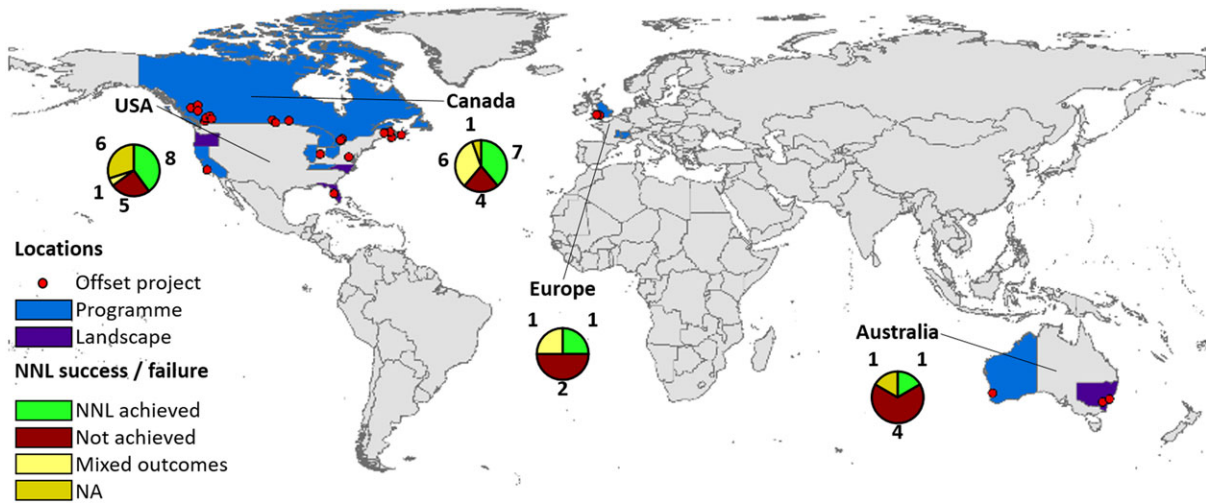


FIGURE 1 Map of all of study and project areas included in our review. Pie-charts indicate the number of projects/studies by region reporting achieving NNL, failing to achieve NNL, achieving a mixture of outcomes for different outcome variables, and for which no NNL designation could be made because the outcome variable was a measure of regulatory compliance

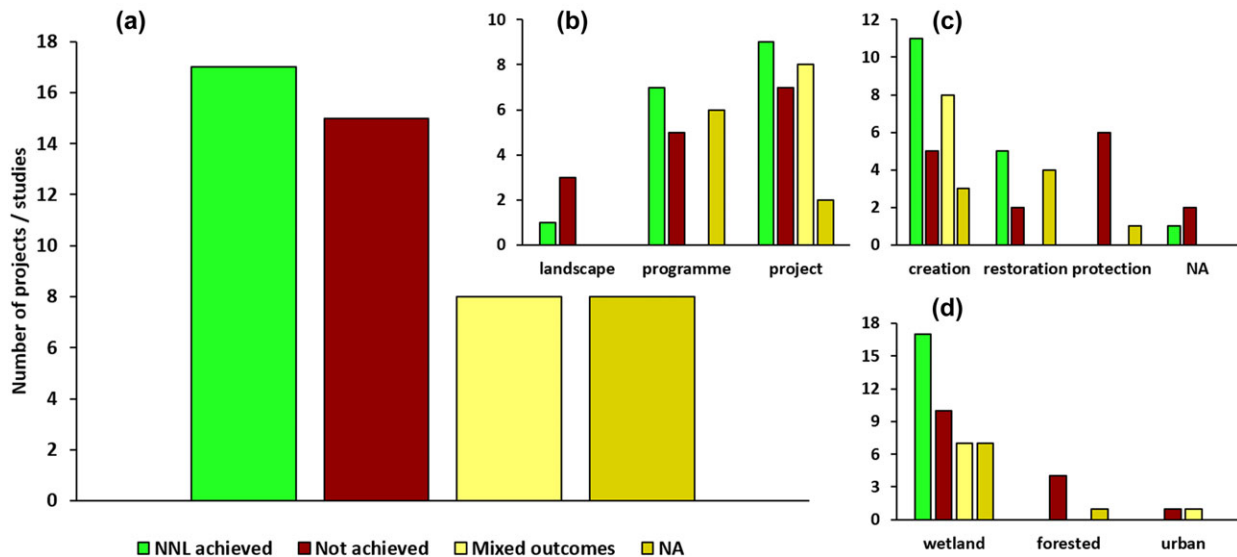


FIGURE 2 (a) Total number of studies/projects within our database achieving NNL. The number of studies/projects is disaggregated by spatial scale (b), offset type (c), and biodiversity type affected (d). NA represents either studies that presented outcome variables from which an NNL designation could not be determined (a), or studies where information on offset type was not provided (c). Studies evaluating the outcomes of bat mitigation actions aiming to achieve NNL in bat population status are categorized as “urban” (d)

ecological outcomes they reported related to whether regulatory compliance standards were met (e.g., percentage invasive species plant cover), which often do not explicitly aim to achieve NNL of biodiversity per se at project scales (Sudol & Ambrose, 2002). When treating each offset or NNL policy independently ($N = 48$), NNL was achieved for 17 assessments, not achieved for 15 assessments, and both successful and unsuccessful depending on the choice of outcome variable for eight assessments. No studies demonstrated the achievement of NNL in forested ecosystems or for avoided loss offsets (Figure 2) (dataset included in Supporting Information).

3.2 | Outcomes of program- and landscape-scale evaluations

Four studies conducted landscape-scale evaluations of the area of land cover changes under the jurisdiction of NNL policies, with three finding that NNL was not achieved by area (Figure 2). No causal interpretation should be given to these results as other conservation policies may have been implemented simultaneously with NNL policies. Levrel, Scemama, and Vaissière (2017) and Carle (2011) focused on Florida and 20 counties across North Carolina, respectively. Both found that total wetland area decreased over their study

periods (2001–2011 and 1994–2001), despite considerable restoration efforts attributable to wetland mitigation policy. Drielsma et al. (2016) evaluated the Southern Mallee Guidelines scheme in western New South Wales, Australia. The authors modeled biodiversity change attributable to the scheme, concluding that it broadly achieved the aim of maintaining or improving native vegetation. However, discounting modeled outcomes, the observed outcomes of the scheme were that over 40,000 ha of vegetated grazing lease were cleared and “offset” through the protection of other areas, leading to an overall net loss in vegetated habitat area. Lastly, Fickas, Cohen, and Yang (2016) found that NNL in wetland area in Willamette Valley (OR) was achieved since the formal adoption of the national No Net Loss policy goal and major clarifications to Section 404 of the Clean Water Act in 1990.

Of the 12 program-scale evaluations in the literature that included outcome variables from which NNL assessments could be made, seven reported achieving NNL (Figure 2). All seven used change in habitat area as outcome variables, and reported results from offset programs focused predominantly on habitat creation and restoration (BenDor, Brozovic, & Pallathucheril, 2007, 2009; Breaux et al., 2005; Harper & Quigley, 2005; Kettlewell et al., 2008; Kozich & Halvorsen, 2012; Robertson & Hayden, 2015). The other three studies also using area as their outcome variables that failed to achieve NNL were all reporting results from offset systems based predominantly on avoided loss offsets (Gibbons et al., 2018; Goldberg & Reiss, 2016; Morgan & Roberts, 2003). The remaining studies evaluated the success of bat mitigation in the United Kingdom under the objective of “NNL in local bat population status,” and the percentage of offset sites in Isère, France, where the required offset habitat type or species was present. Here, NNL was not achieved for both bat presence and abundances post-mitigation (categorized as “urban” in Figure 2; Stone, Jones, & Harris, 2013), and offset habitat/species presence varied from 61% to 73% (Bezombes, Kerbiriou, & Spiegelberger, 2019).

3.3 | Outcomes of biodiversity offsets

Twenty-six biodiversity offsets from 10 studies were included in our database, of which we could make NNL designations for 24. Of these, nine achieved NNL for all given outcome variables, seven failed to achieve any, and eight achieved NNL for some outcome variables but not for others (Figure 3). There was not enough identifying variation in the data to statistically explore whether specific aspects of offset design, type, or ecology predicted the achievement of a higher percentage of total outcome variables. Nevertheless, it is noteworthy that 64% (7/11, Figure 3c) of projects with offset ratios >1 achieved NNL for all of their associated outcome variables compared with 17% for offsets with ratios ≤1 (2/12, Figure 3b).

There was nominally variation between outcome measures when comparing outcome values between offset and impact/reference sites (Figure 4), although an insufficient data volume to explore statistical differences. On average, assessments of habitat quality tended to find that the quality of offset sites was lower than that at impact sites.

For the eight project-based studies where offsets were ecologically compared with either their impact sites or reference sites, three met all our criteria for study validity (Garland, Wells, & Markham, 2017; Teels, Mazanti, & Rewa, 2004; Thorn et al., 2018). Two sampled control/offset sites at a single time-point and thus were unable to account for natural ecological variability in outcomes (Hegberg, Baker, & Pieper, 2010; Quigley & Harper, 2006a, but see justification in Quigley & Harper, 2006a), one did not report its sampling protocol and is thus open to sampling bias (Hegberg et al., 2010), and four used controls for their NNL assessments, which were collected ≥5 years before data at the offset site (Hegberg et al., 2010; Lindenmayer et al., 2017; Murata & Feest, 2015; Pickett et al., 2013).

3.4 | Outcomes of studies evaluating compliance

Ten studies evaluated the degree to which NNL implementation was meeting regulatory compliance standards at programme scales (Figure 5). Compliance across NNL programmes was imperfect, with no compliance rates exceeding 75% (Hill, Kulz, Munoz, & Dorney, 2013).

3.5 | Reasons for NNL achievement or failure

The two most commonly cited reasons for a lack of NNL success were: failure of the specific conservation interventions applied by the offset (e.g., the offset species failing to respond as expected to the offsetting intervention); and offset implementation failures (Table 3). The most commonly cited reason for success was having high offset ratios. Additionally, Fickas et al. (2016) noted that NNL policy internalized impacts on wetlands that were previously not subject to regulation, thus potentially disincentivizing habitat conversion.

4 | DISCUSSION

Our review reveals important insights about the state of the evidence base for NNL and biodiversity offsetting. We provide preliminary indications that: NNL has historically been more successful in wetland than forested ecosystems; avoided loss offsets are particularly risky; evaluations have so far predominantly used area-based outcome measures; there are potential problems with the validity of studies evaluating

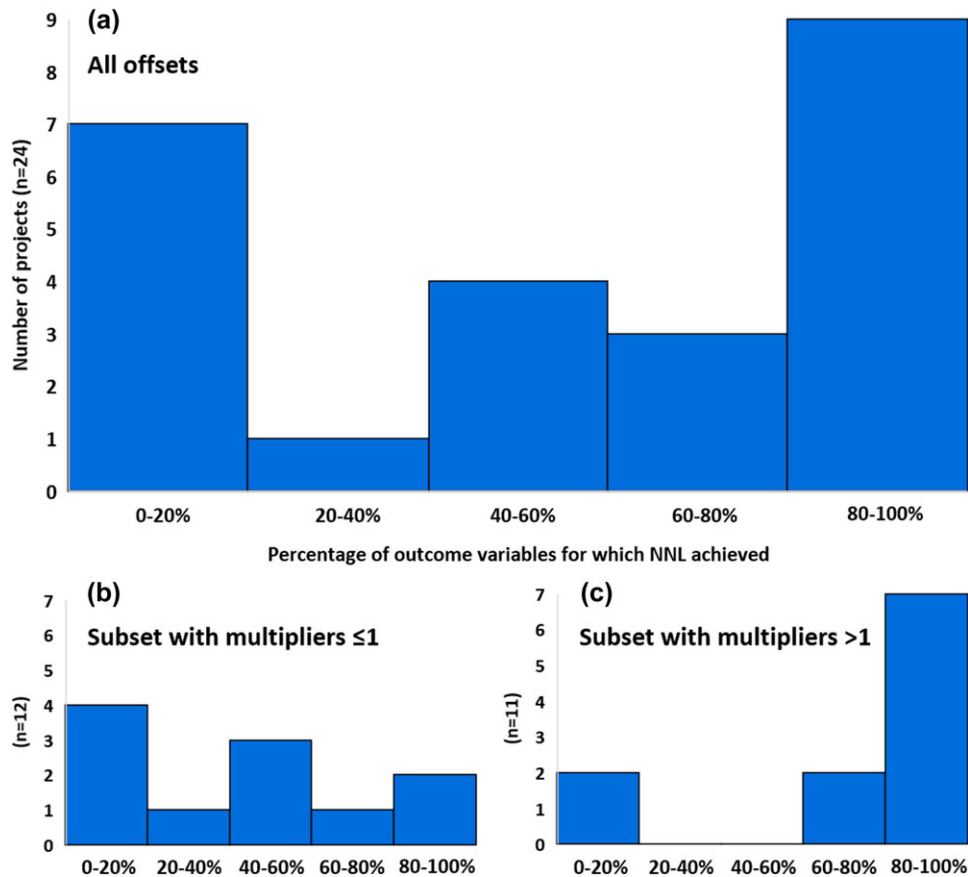


FIGURE 3 (a) Frequency distribution of the percentage of outcome variables achieved for each offset project in our sample where an NNL designation could be made (including one avoided loss offset that is excluded from (b) and (c) (Thorn et al., 2018)). (b) For all creation/restoration offset projects with a multiplier ≤ 1 . (c) For all creation/restoration projects with a multiplier > 1

offset outcomes; and the most common reason for offset success appears to be the implementation of high offset ratios.

We identify a substantial gap between the global implementation of NNL and the evidence base concerning ecological effectiveness. Sixty-seven percent of the world's offsets are applied in forested ecosystems (Bull & Strange, 2018), yet our review reveals that only four studies have assessed NNL outcomes from offsets applied to forest ecosystems or wildlife. Of these, none demonstrated that their associated NNL targets were achieved. Similarly, 20% of the world's offsets entail some form of protection or avoided loss (Bull & Strange, 2018). Yet, only six studies have assessed NNL outcomes from this common offset type, and none found that NNL was achieved.

4.1 | Exploring unsuccessful outcomes of avoided loss and forest offsets

Avoided loss offsets appeared to be unsuccessful for multiple reasons. Critically, they necessarily lead to an immediate net loss in habitat area (Gibbons et al., 2018). This can be justified as a mechanism for preventing biodiversity loss if the background rate of biodiversity loss is sufficiently high. How-

ever, in the studies included here and the wider literature, it is evident that assumed rates of background declines are commonly higher than the actual rate, superficially justifying the use of avoided loss offsets when in reality gains only accrue many decades into the future (Gibbons et al., 2018; Reside et al., 2019). This issue is compounded if the "protection" afforded by offsets does not actually reduce the probability of loss, most commonly when sites that are not under threat of development receive "protection" (e.g., Thorn et al., 2018). Drielsma et al. (2016) justify the use of avoided loss on the grounds that biodiversity improvements on newly protected sites could offset the losses attributable to the reduction in overall habitat extent. Whether these condition gains are achieved in reality is questionable, especially considering the consistent ecological or implementation failure of conservation management interventions associated with offsets in our sample (Bezombes et al., 2019; Lindenmayer et al., 2017).

Many of the same reasons apply to explain the apparent failure of offsets focused on forest biodiversity, although identifying explanations unique to forests is challenging as four of five forest offset studies are also avoided loss offsets. Additionally, all forest studies came from Australia, where native vegetation offsets based predominantly on avoided loss have been

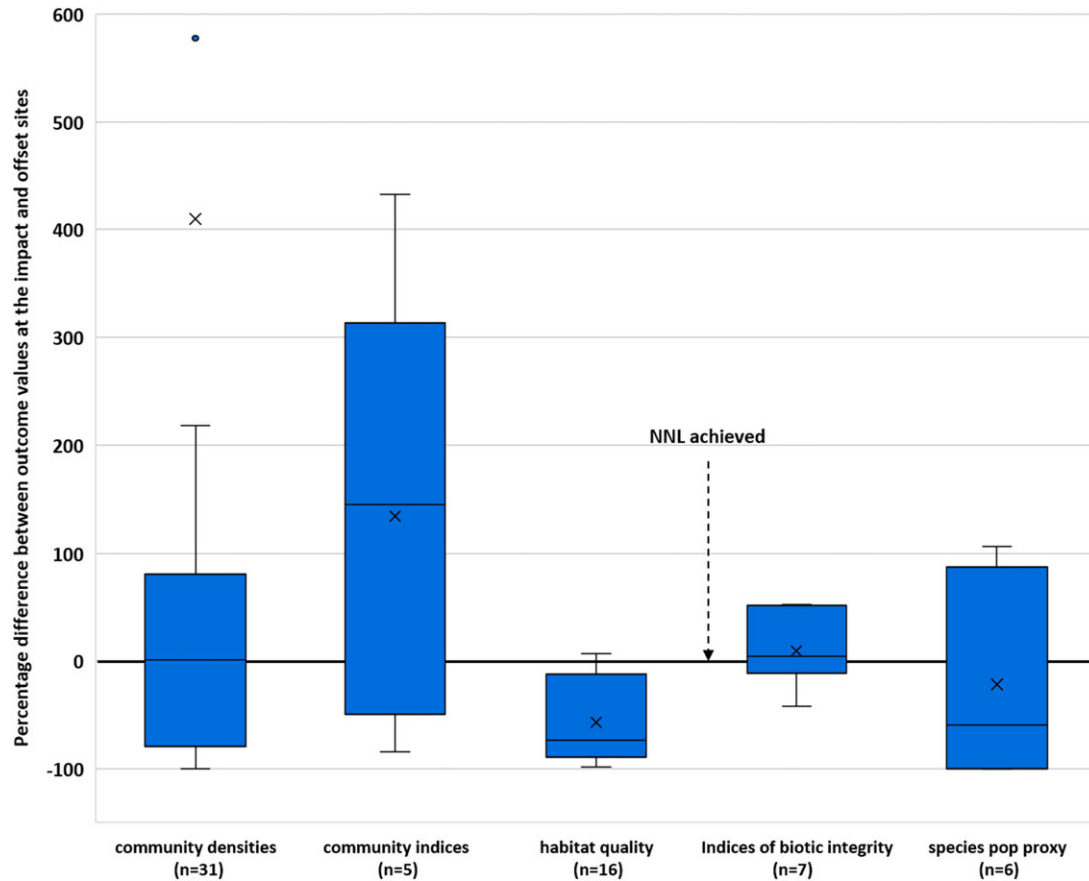


FIGURE 4 Box and whisker plots showing the upper and lower quartiles and exclusive medians of the percentage difference between outcome values at offset sites relative to impact/control sites, with outcome variables grouped into categories. Whiskers indicate the maximum/minimum values that fall within $\pm 1.5 \times$ inter-quartile range. Values > 0 indicate that the value at the offset site exceeded that at the impact site. Four outliers (represented by dots) not shown: for the “community densities” column, outliers occurred at 1469, 3093, 3426, and 4348. Outliers are likely explained by Quigley and Harper (2006a) containing several projects with unusually high offset ratios at several of the sites, and the use of stochastic community density-based outcome measures (e.g., number of invertebrates sampled/m²). Crosses denote the sample mean. See Table S3 for summary of which outcome variables were assigned to each category

criticized for facilitating high rates of deforestation and species declines (Reside et al., 2019). Nevertheless, both studies evaluating interventions aiming to offset impacts on forest species found that the interventions failed to deliver ecological equivalence, providing either lower quality or less-utilized habitat than that impacted by development (Lindenmayer et al., 2017; Thorn et al., 2018). On the planning side, May, Hobbs, and Valentine (2017) identify a number of shortcomings hindering Western Australia’s native vegetation offset policies from achieving NNL, including a lack of contingency planning in the case of offset failure, insufficient reporting of offset outcomes, offset performance criteria being disconnected from actual ecological outcomes, and poor compliance.

4.2 | Compliance with NNL policies

May et al.’s (2017) findings are indicative of the rest of the evaluations of compliance in our dataset, with variously

defined compliance rates ranging from 4% to 75%. Imperfect compliance rates per se do not guarantee failure of NNL policies from an ecological perspective, as the effects of compliance failure might be outweighed by offset multipliers (Bull, Lloyd, & Strange, 2017). However, a recent global review including gray literature demonstrated that compliance with offset permit criteria often considerably exceeds the ecological functional performance of those offsets, indicating that achieving compliance is often insufficient to achieve NNL (Theis et al., 2019). Additionally, low compliance rates do indicate that regulatory enforcement of offset outcomes is often lacking, potentially demonstrating limited institutional interest in the true outcomes of offsetting, thus weakening the probability of NNL outcomes (Walker et al., 2009). There are rarely legal mechanisms for imposing financial penalties for non-compliance (Hahn & Richards, 2013). Improving monitoring alone will not guarantee improved outcomes (Kozich & Halvorsen, 2012): compliance likely requires strict enforcement, with regulators empowered to impose

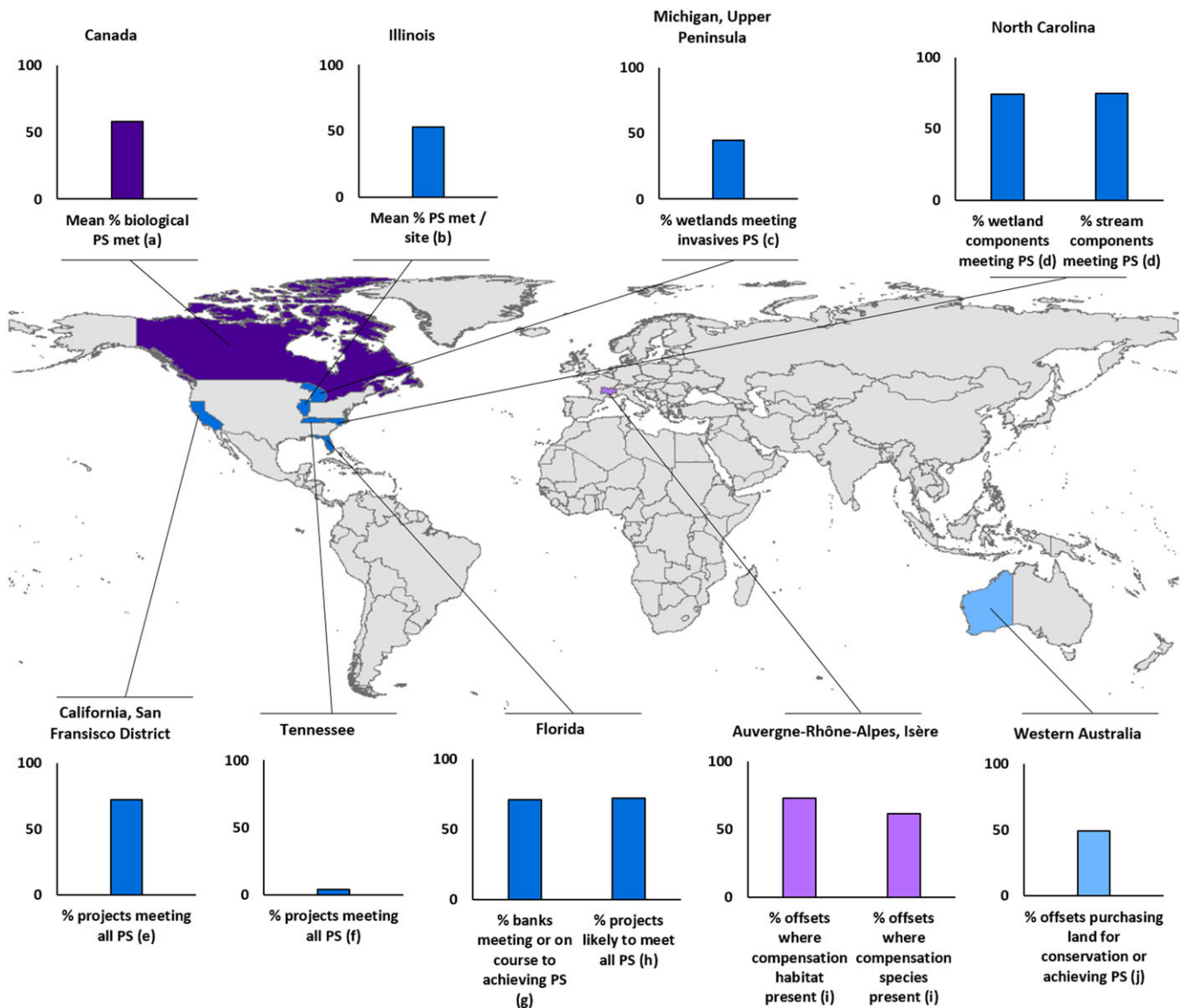


FIGURE 5 Percentage compliance and compliance criteria reported for regions in our dataset, with bar chart colors corresponding to the region providing the compliance values. Note that the type of reported compliance standards varies between studies, so rates are not comparable. PS denotes “performance standards”. (a) Quigley and Harper (2006b); (b and i) Bezombes et al. (2019); (e) Breaux et al. (2005); (d) Hill et al. (2013); (c) Kozich and Halvorsen (2012); (j) May et al. (2017); (f) Morgan and Roberts (2003); (g) Reiss, Hernandez, and Brown (2009); (h) Shafer and Roberts (2008); Van den Bosch and Matthews (2017)

punishments when permits are violated (Gray & Shimshack, 2011). Such pecuniary enforcement measures have been demonstrated in the context of other environmental policies to have direct and indirect benefits, such as both increasing compliance rates within punished firms, and inducing spillovers improving compliance within unpunished firms (Gray & Shimshack, 2011). While improving compliance is likely key, if NNL policies fail to use an appropriate reference system (either the pre-impact site or control site) to define the compliance criteria for offsets, then even achieving full compliance may well fail to achieve NNL of biodiversity across the paired impacted and offset sites (Theis et al., 2019).

4.3 | Achieving NNL: True success or methodological artifact?

Despite little evidence for the effectiveness of some common offset types, one-third of all projects or studies in our database reported achieving NNL. All but one of the successful NNL outcomes occurred for wetland habitats or species, with 50% of wetland projects/studies where an NNL designation could be made achieving NNL. Additionally, all of the successful NNL outcomes occurred for creation or restoration offsets. We speculate that wetland restoration offsets might have higher NNL rates than other offset types in our dataset for two main reasons: first, wetlands display

TABLE 3 List of reasons cited for NNL policy/offset success or failure. The number of citations per reason should not be taken to indicate the importance of that reason, as there was variation between papers in the depth of their discussions of potential explanations

NNL outcome	Reason	Scale	References
NNL/offset <u>failure</u> , failure to achieve compliance	Avoided loss leading to an overall loss in area of natural habitats	Program	Morgan and Roberts, 2003; Gibbons et al., 2018
	Compliance standards unrelated to ecological outcomes	Program	May et al., 2017
	Conflict with development	Program	Shafer and Roberts, 2008
	Conservation intervention failure	Program; project	Quigley and Harper, 2006b; Stone et al., 2013; Lindenmayer et al., 2017; Garland et al., 2017; Bezombes et al., 2019
	Contradictions within permit requirements	Program	Quigley and Harper, 2006b
	Failure to consider landscape context	Program	Van den Bosch and Matthews, 2017
	Illegal trespassing	Program	Hill et al., 2013
	Insufficient offset ratios	Program; project	Goldberg and Reiss, 2016; Stone et al., 2013; Quigley and Harper, 2006a
	Invasive encroachment without management	Program	Van den Bosch and Matthews, 2017
	Lack of additionality	Project	Thorn et al., 2018
	Lack of contingency measures in case of offset failure	Program	May et al., 2017
	Lack of data to demonstrate outcomes	Program	May et al., 2017; Stone et al., 2013; Bezombes et al., 2019
	Lack of ecological equivalence	Project	Thorn et al., 2018; Teels et al., 2004
	Lack of ecological suitability of creation offset site	Program	Hill et al., 2013; Kozich and Halvorsen, 2012; Quigley and Harper, 2006a; Shafer and Roberts, 2008
	Lack of monitoring	Program; project	Quigley and Harper, 2006a; Quigley and Harper, 2006b
	Lack of offset expertise	Program	Quigley and Harper, 2006b
	Offset implementation failure	Program	Quigley and Harper, 2006b; May et al., 2017; Morgan and Roberts, 2003; Shafer and Roberts, 2008; Bezombes et al., 2019
	Temporal lag	Program; project	Quigley and Harper, 2006a
	Unregulated impacts	Landscape; Program	Goldberg and Reiss, 2016; Carle, 2011
NNL/offset <u>success</u>	Bringing impacts under regulation	Landscape	Fickas et al., 2016
	High offset ratio	Program; project	Pickett et al., 2013; Robertson and Hayden, 2015; Harper and Quigley, 2005
	Simple biodiversity metric	Project	Pickett et al., 2013

higher rates of ecological recovery than many other habitat types (Jones et al., 2018), and this recovery is more likely to reach reference conditions if the impacted wetland was itself degraded (relatively likely in areas undergoing development or construction). Second, the two main wetland offsetting policies covered by our dataset are Section 404 of the Clean Water Act in the USA and the Canadian policy of NNL in productive fish habitat. These rank among the oldest NNL policies, and both have undergone numerous refinements during their implementation (Hough & Robertson, 2009; Rubec & Hanson, 2009), thus, their effectiveness might exceed that of younger offset policies elsewhere.

An additional key reason for biodiversity offset success appears to be high offset ratios. This finding should be considered in the context of recent literature encouraging practitioners not to simply rely upon high multipliers to solve all offset implementation problems (Bull et al., 2017). However, within our database, high multipliers appear to be a predictor of NNL success. For individual species-based offsets, this may be because high offset multipliers can be a useful mechanism for increasing habitat availability for the offset species and thus easing density-dependence constraints within the re-establishing population (Pickett et al., 2013). For habitat-based offsets, high multipliers might promote

the achievement of NNL if best-practice biodiversity metrics which account for both habitat extent and condition are used (Bezombes et al., 2018), although care must be taken to constrain trades between habitat condition types to avoid trading large extents of biodiversity-poor habitat for small extents of valuable habitat (Carver & Sullivan, 2017).

However, it is unclear to what degree these perceived predictors of success (wetlands and high multipliers) reflect true trends, or whether these reflect the choice of outcome variables used to assess NNL. At program scales, seven of 10 wetland studies where a NNL designation could be made found that NNL was achieved, but all studies used area as an outcome variable. At landscape scales, two of three wetland studies found that NNL was not achieved, and again all used area as their outcome variables. At project scales, nine of 21 offsets achieved NNL, yet for seven of these successes the outcome variables were community densities. Six of these successes came from Quigley and Harper (2006a), who calculated whether NNL was achieved for community density outcomes while accounting for the offset multiplier (i.e., to infer whether the overall abundance of the community group in question was higher for the offset than the impact site). Thus, these successful NNL outcomes are also linked inextricably to offset area. Therefore, with our current dataset we cannot definitively answer the question of whether true NNL in biodiversity is more likely for wetlands than other habitat types, because many of the current metrics used to assess NNL in the literature are confounding offset area (and the offset multiplier) with increases in biodiversity. This is problematic because habitat area alone does not necessarily reflect habitat quality or community composition (Dale & Gerlak, 2007), and is thus widely recognized as an unsatisfactory biodiversity metric (Quétier & Lavorel, 2011). Additionally, this review cannot indicate the direction of causality—projects with larger offset ratios might be more likely to be successful, but plausibly larger offset ratios might merely be more strongly embedded into older NNL policies.

4.4 | Influence of spatial scale on NNL outcomes

The perceived discrepancy in outcomes between landscape-scale and program-scale evaluations of NNL is likely because program-scale evaluations only account for registered offsets/impacts: yet unregulated or exempt impacts may well make the difference between achieving NNL or not (Maron et al., 2018). For example, in Florida between 2001 and 2011, mitigation banking restored 58,575 ha of wetlands, yet across the state, a net 5600 ha/year was lost during the same time period (Levrel et al., 2017), which is possibly because the Clean Water Act applies only to “jurisdictional wetlands,” thus many wetland impacts escape regulation. Discrepancies between the apparent success of program-scale area-based

evaluations and landscape-scale ones indicate that NNL policies are likely undermined if some impacts are unreported or otherwise exempt from regulation (Gibbons et al., 2018; Reside et al., 2019). Thus, the scope of impacts falling underneath these policies should be widened to include all impacts and minimize opportunities to avoid NNL legislation.

4.5 | Outcomes of individual biodiversity offsets

For individual offsets, the outcome variables used were more complex than merely habitat area, and generally adapted to the particular contexts of their associated NNL policies (e.g., Quigley & Harper, 2006a used indicators of habitat productivity, variables representing habitat quality and community densities, to assess whether offsets achieved their policy target of NNL of productive fish habitat). Notably, we found only three studies that attempted to assess whether offset and impact sites were ecologically equivalent at the community level. For offsetting to be demonstrably ecologically equivalent, it should capture aspects of species identity or community composition: two studies accounted for community composition by using indices of biotic integrity (Hegberg et al., 2010; Teels et al., 2004), and one by assessing whether habitat type, quality, and structure were similar to that at the impact site (Thorn et al., 2018). Given the strong emphasis in best-practice principles on achieving ecological equivalence (McKenney & Kiesecker, 2010; Quétier & Lavorel, 2011), the lack of empirical evaluations demonstrating equivalence is a clear gap.

Additionally, we found a number of methodological issues with offset studies, with three of eight studies conducting site-based ecological assessments of biodiversity losses and gains meeting our criteria for study internal validity. Alongside opportunities for selection bias and the measurement of biodiversity at a single timepoint that does therefore not account for ecological dynamics, the most common issue was the use of controls that are open to potential performance bias (Bilotta et al., 2014). Four studies used controls from ≥ 5 years before measuring biodiversity at the offset site, which can be justified on the grounds that development projects take years to be implemented, but it cannot be ruled out that other factors influenced changes in biodiversity over this time, thus obscuring the true impact of the NNL policy on biodiversity. Additionally, although not identified in these studies, evaluators should beware pseudoreplication when assessing whether NNL is achieved across multiple sites.

Combined, these points emphasise the need for higher-quality evidence to understand when NNL is defensible as a conservation strategy. Our review identified just one study meeting our inclusion criteria that compared NNL outcomes with a robust counterfactual (Gibbons et al., 2018). Generally, the quality of impact evaluations for NNL appear to be lagging behind those applied in other areas of conservation and

environmental policy, such as payments for ecosystem services (Pynegar, Jones, Gibbons, & Asquith, 2018), protected areas (Miteva, Pattanayak, & Ferraro, 2012), commodity sustainability certification (Carlson et al., 2018), and forest policy (Simmons et al., 2018). Recognizing that the true causal impact of conservation policies can be confounded by biases in those receiving conservation treatments, there is an increase in applications of experimental, quasi-experimental, and matching methods to improve our causal understanding of policy effectiveness (Ferraro & Hanauer, 2014). The first study of this kind assessing the effectiveness of NNL-related policies focused on species conservation banks (Sonter, Barnes, Matthews, & Maron, 2019), but most of these do not have NNL requirements, and to date, there remain no NNL evaluations using advanced causal inference. This is, therefore, a vital area of future research.

Biodiversity offsets receive disproportionate attention compared to the other stages of the MH (Hough & Robertson, 2009). However, the effectiveness of NNL is fundamentally reliant on robust implementation of avoidance and minimization measures (Phalan et al., 2018; von Hase & ten Kate, 2017). Our current understanding of the effectiveness of these stages is limited. The major difficulty in evaluating avoidance is that only part of the process of avoidance is observable: permit denials and evaluations of alternative impact sites common to major infrastructure projects. The evidence from these stages would imply that avoidance is weakly applied, as numerous studies have demonstrated low rates of project rejection on environmental grounds and weak justifications for why final project sites were chosen (Clare, Krogman, Foote, & Lemphers, 2011; Phalan et al., 2018). However, recent work from South Africa has found these observable characteristics to be imperfect reflections of the actual avoidance embedded in the planning process, as many decisions on avoidance happen through informal consultations with regulators in advance of project proposal (Sinclair, 2018).

4.6 | Policy implications

Finally, are the findings of this review generalizable and of policy relevance? Our search language is a limitation, and while there is evidence that English captures most of the literature on NNL implementation globally (Bull & Strange, 2018), NNL systems in countries such as Germany or Brazil may not have been captured in our review. Furthermore, our sampling strategy is biased away from the gray literature. However, the direction of this bias is unclear (Theis et al., 2019)—plausible arguments could be made both for a selection bias toward publishing unsuccessful NNL results in the academic conservation literature, and toward not publishing unsuccessful results in the gray literature because of a fear of criticism for legislators or vested interests. Additionally, although our review was global, the evaluations of actual

NNL outcomes identified in our review are biased toward high-income countries with strong institutions. Thus, it is possible that our review may overestimate the probability of achieving NNL outcomes in countries with weaker environmental legislation. However, strong institutions far from guarantee a successful NNL policy—details of NNL design are vitally important (Maron et al., 2018). Therefore, without overstating our findings, we feel there are generalizable recommendations that can be derived from our review:


- policymakers should be aware that without significant improvements to existing policies, NNL policies in forested habitats or utilizing avoided loss offsets are unlikely to achieve NNL;
- improving compliance with NNL policies is essential for achieving improved ecological outcomes (which may come from mandating some form of penalty for noncompliance);
- and it is important to move beyond area-based outcome measures when implementing NNL.

With \$60–70 trillion dollars committed to infrastructural expansion by 2030 (Laurance et al., 2015), it is essential that we develop solutions that fully address the unmitigated biodiversity impacts of infrastructural expansion. If we are to achieve NNL of biodiversity, it is an urgent priority to develop the evidence base to understand what works, and when.


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REFERENCES

- BenDor, T., Brozovic, N., & Pallathucherial, V. G. (2007). Assessing the socioeconomic impacts of wetland mitigation in the Chicago region. *Journal of the American Planning Association*, 73, 263–282.
- BenDor, T., Sholtes, J., & Doyle, M. W. (2009). Landscape characteristics of a stream and wetland mitigation banking program. *Ecological Applications*, 19, 2078–2092.

- Bennett, G., Gallant, M., & ten Kate, K. (2017). *State of biodiversity mitigation 2017: Markets and compensation for global infrastructure development*. Washington, DC: Forest Trends' Ecosystem Marketplace.
- Bezombes, L., Gaucherand, S., Spiegelberger, T., Gouraud, V., & Kerbiriou, C. (2018). A set of organized indicators to conciliate scientific knowledge, offset policies requirements and operational constraints in the context of biodiversity offsets. *Ecological Indicators*, *93*, 1244–1252.
- Bezombes, L., Kerbiriou, C., & Spiegelberger, T. (2019). Do biodiversity offsets achieve No Net Loss? An evaluation of offsets in a French department. *Biological Conservation*, *231*, 24–29.
- Bilotta, G. S., Milner, A. M., & Boyd, I. L. (2014). Quality assessment tools for evidence from environmental science. *Environmental Evidence*, *3*, 14.
- Breaux, A., Cochrane, S., Evens, J., Martindale, M., Pavlik, B., Suer, L., & Benner, D. (2005). Wetland ecological and compliance assessments in the San Francisco Bay Region, California, USA. *Journal of Environmental Management*, *74*, 217–237.
- Bull, J., Gordon, A., Law, E., Suttle, K., & Milner-Gulland, E. (2014). Importance of baseline specification in evaluating conservation interventions and achieving no net loss of biodiversity. *Conservation Biology*, *28*, 799–809.
- Bull, J. W., Brauneder, K., Darbi, M., Van Teeffelen, A. J., Quétier, F., Brooks, S. E., ... Strange, N. (2018). Data transparency regarding the implementation of European 'no net loss' biodiversity policies. *Biological Conservation*, *218*, 64–72.
- Bull, J. W., Lloyd, S. P., & Strange, N. (2017). Implementation gap between the theory and practice of biodiversity offset multipliers. *Conservation Letters*, *10*, 656–669.
- Bull, J. W., & Strange, N. (2018). The global extent of biodiversity offset implementation under no net loss policies. *Nature Sustainability*, *1*, 790–798.
- Bull, J. W., Suttle, K. B., Gordon, A., Singh, N. J., & Milner-Gulland, E. (2013). Biodiversity offsets in theory and practice. *Oryx*, *47*, 369–380.
- Calvet, C., Napoléone, C., & Salles, J.-M. (2015). The biodiversity offsetting dilemma: Between economic rationales and ecological dynamics. *Sustainability*, *7*, 7357–7378.
- Carle, M. V. (2011). Estimating wetland losses and gains in coastal North Carolina: 1994–2001. *Wetlands*, *31*, 1275–1285.
- Carlson, K. M., Heilmayr, R., Gibbs, H. K., Noojipady, P., Burns, D. N., Morton, D. C., ... Kremen, C. (2018). Effect of oil palm sustainability certification on deforestation and fire in Indonesia. *Proceedings of the National Academy of Sciences of the United States of America*, *115*, 121–126.
- Carver, L., & Sullivan, S. (2017). How economic contexts shape calculations of yield in biodiversity offsetting. *Conservation Biology*, *31*, 1053–1065.
- Clare, S., Krogman, N., Foote, L., & Lemphers, N. (2011). Where is the avoidance in the implementation of wetland law and policy? *Wetlands Ecology and Management*, *19*, 165–182.
- Dale, L., & Gerlak, A. K. (2007). It's all in the numbers: Acreage tallies and environmental program evaluation. *Environmental Management*, *39*, 246–260.
- Drielsma, M. J., Foster, E., Ellis, M., Gill, R. A., Prior, J., Kumar, L., ... Ferrier, S. (2016). Assessing collaborative, privately managed biodiversity conservation derived from an offsets program: Lessons from the Southern Mallee of New South Wales, Australia. *Land Use Policy*, *59*, 59–70.
- EPA. (2008). Compensatory mitigation for losses of aquatic resources, final rule [WWW Document]. Retrieved from https://www.epa.gov/sites/production/files/2015-03/documents/2008_04_10_wetlands_wetlands_mitigation_final_rule_4_10_08.pdf
- Ferraro, P. J., & Hanauer, M. M. (2014). Advances in measuring the environmental and social impacts of environmental programs. *Annual Review of Environment and Resources*, *39*, 495–517.
- Fickas, K. C., Cohen, W. B., & Yang, Z. (2016). Landsat-based monitoring of annual wetland change in the Willamette Valley of Oregon, USA from 1972 to 2012. *Wetlands Ecology and Management*, *24*, 73–92.
- Garland, L., Wells, M., & Markham, S. (2017). Performance of artificial maternity bat roost structures near Bath, UK. *Conservation Evidence*, *14*, 44–51.
- Gibbons, P., Macintosh, A., Constable, A. L., & Hayashi, K. (2018). Outcomes from 10 years of biodiversity offsetting. *Global Change Biology*, *24*, e643–e654.
- Goldberg, N., & Reiss, K. C. (2016). Accounting for wetland loss: Wetland mitigation trends in northeast Florida 2006–2013. *Wetlands*, *36*, 373–384.
- Gordon, A., Bull, J. W., Wilcox, C., & Maron, M. (2015). Perverse incentives risk undermining biodiversity offset policies. *Journal of Applied Ecology*, *52*, 532–537.
- Gray, W. B., & Shimshack, J. P. (2011). The effectiveness of environmental monitoring and enforcement: A review of the empirical evidence. *Review of Environmental Economics and Policy*, *5*, 3–24.
- Hahn, R., & Richards, K. (2013). Understanding the effectiveness of environmental offset policies. *Journal of Regulatory Economics*, *44*, 103–119.
- Harper, D., & Quigley, J. (2005). No net loss of fish habitat: A review and analysis of habitat compensation in Canada. *Environmental Management*, *36*, 343–355.
- von Hase, A., & ten Kate, K. (2017). Correct framing of biodiversity offsets and conservation: A response to Apostolopoulou & Adams. *Oryx*, *51*, 32–34.
- Hegberg, C., Baker, A., & Pieper, M. (2010). Stream relocation design and monitoring in a rapidly urbanizing watershed: EFWLC, Indianapolis, Indiana. In D. K. Frevert & K. W. Potter (Eds.), *Watershed Management 2010: Innovations in watershed management under land use and climate change* (Vol. 394, pp. 457–468). Reston, VA: American Society of Civil Engineers.
- Hill, T., Kulz, E., Munoz, B., & Dorney, J. R. (2013). Compensatory stream and wetland mitigation in North Carolina: An evaluation of regulatory success. *Environmental Management*, *51*, 1077–1091.
- Hobbs, R., & MacAller, R. (2014). Santa Maria River levee ecological restoration program. Presented at the Legislation, Technology and Practice of Mine Land Reclamation: Proceedings of the Beijing International Symposium on Land Reclamation and Ecological Restoration (LRER 2014), Beijing, China, 16–19 October 2014, CRC Press, p. 199.
- Hough, P., & Robertson, M. (2009). Mitigation under Section 404 of the Clean Water Act: Where it comes from, what it means. *Wetlands Ecology and Management*, *17*, 15–33.
- IPBES. (2019). Intergovernmental Platform on Biodiversity and Ecosystem Services - Summary for policymakers of the global assessment report [WWW Document]. Retrieved from https://www.ipbes.net/sites/default/files/downloads/spm_unedited_advance_for_posting_htn.pdf

- Ives, C. D., & Bekessy, S. A. (2015). The ethics of offsetting nature. *Frontiers in Ecology and the Environment*, *13*, 568–573.
- Jones, H. P., Jones, P. C., Barbier, E. B., Blackburn, R. C., Rey Benayas, J. M., Holl, K. D., ... Mateos, D. M. (2018). Restoration and repair of Earth's damaged ecosystems. *Proceedings of the Royal Society: B, Biological sciences*, *285*, 20172577.
- Karr, J. R. (1981). Assessment of biotic integrity using fish communities. *Fisheries*, *6*, 21–27.
- Kettlewell, C. I., Bouchard, V., Porej, D., Micacchion, M., Mack, J. J., White, D., & Fay, L. (2008). An assessment of wetland impacts and compensatory mitigation in the Cuyahoga River Watershed, Ohio, USA. *Wetlands*, *28*, 57–67.
- Khangura, S., Konnyu, K., Cushman, R., Grimshaw, J., & Moher, D. (2012). Evidence summaries: The evolution of a rapid review approach. *Syst. Rev.*, *1*, 10.
- Kozich, A. T., & Halvorsen, K. E. (2012). Compliance with wetland mitigation standards in the Upper Peninsula of Michigan, USA. *Environmental Management*, *50*, 97–105.
- Lajeunesse, M. J. (2016). Facilitating systematic reviews, data extraction and meta-analysis with the metagear package for R. *Methods in Ecology and Evolution*, *7*, 323–330.
- Laurance, W. F., Peletier-Jellema, A., Geenen, B., Koster, H., Verweij, P., Van Dijck, P., ... Van Kuijk, M. (2015). Reducing the global environmental impacts of rapid infrastructure expansion. *Current Biology*, *25*, R259–R262.
- Levrel, H., Scemama, P., & Vaissière, A.-C. (2017). Should we be wary of mitigation banking? Evidence regarding the risks associated with this wetland offset arrangement in Florida. *Ecological Economics*, *135*, 136–149.
- Lindenmayer, D. B., Crane, M., Evans, M. C., Maron, M., Gibbons, P., Bekessy, S., & Blanchard, W. (2017). The anatomy of a failed offset. *Biological Conservation*, *210*, 286–292.
- Maron, M., Brownlie, S., Bull, J. W., Evans, M. C., von Hase, A., Quétier, F., ... Gordon, A. (2018). The many meanings of no net loss in environmental policy. *Nature Sustainability*, *1*, 19–27.
- Maron, M., Hobbs, R. J., Moilanen, A., Matthews, J. W., Christie, K., Gardner, T. A., ... McAlpine, C. A. (2012). Faustian bargains? Restoration realities in the context of biodiversity offset policies. *Biological Conservation*, *155*, 141–148.
- Maxwell, S. L., Fuller, R. A., Brooks, T. M., & Watson, J. E. (2016). Biodiversity: The ravages of guns, nets and bulldozers. *Nature*, *536*, 143–145.
- May, J., Hobbs, R. J., & Valentine, L. E. (2017). Are offsets effective? An evaluation of recent environmental offsets in Western Australia. *Biological Conservation*, *206*, 249–257.
- McKenney, B. A., & Kiesecker, J. M. (2010). Policy development for biodiversity offsets: A review of offset frameworks. *Environmental Management*, *45*, 165–176.
- Miteva, D. A., Pattanayak, S. K., & Ferraro, P. J. (2012). Evaluation of biodiversity policy instruments: What works and what doesn't? *Oxford Review of Economic Policy*, *28*, 69–92.
- Morgan, K. L., & Roberts, T. H. (2003). Characterization of wetland mitigation projects in Tennessee, USA. *Wetlands*, *23*, 65–69.
- Murata, N., & Feest, A. (2015). A case study of evidence for showing 'no net loss' of bird biodiversity in a development project. *Water Environment Journal*, *29*, 419–429.
- Phalan, B., Hayes, G., Brooks, S., Marsh, D., Howard, P., Costelloe, B., ... Whitaker, S. (2018). Avoiding impacts on biodiversity through strengthening the first stage of the mitigation hierarchy. *Oryx*, *52*, 316–324.
- Pickett, E. J., Stockwell, M. P., Bower, D. S., Garnham, J. I., Pollard, C. J., Clulow, J., & Mahony, M. J. (2013). Achieving no net loss in habitat offset of a threatened frog required high offset ratio and intensive monitoring. *Biological Conservation*, *157*, 156–162.
- Pynegar, E. L., Jones, J. P., Gibbons, J. M., & Asquith, N. M. (2018). The effectiveness of Payments for Ecosystem Services at delivering improvements in water quality: Lessons for experiments at the landscape scale. *PeerJ*, *6*, e5753.
- Quétier, F., & Lavorel, S. (2011). Assessing ecological equivalence in biodiversity offset schemes: Key issues and solutions. *Biological Conservation*, *144*, 2991–2999.
- Quigley, J. T., & Harper, D. J. (2006a). Effectiveness of fish habitat compensation in Canada in achieving no net loss. *Environmental Management*, *37*, 351–366.
- Quigley, J. T., & Harper, D. J. (2006b). Compliance with Canada's Fisheries Act: A field audit of habitat compensation projects. *Environmental Management*, *37*, 336–350.
- Reiss, K. C., Hernandez, E., & Brown, M. T. (2009). Evaluation of permit success in wetland mitigation banking: A Florida case study. *Wetlands*, *29*, 907.
- Reside, A. E., Cosgrove, A. J., Pointon, R., Trezise, J., Watson, J. E., & Maron, M. (2019). How to send a finch extinct. *Environmental Science & Policy*, *94*, 163–173.
- Rohatgi, A. (2015). *WebPlotDigitizer*. Austin, Texas, USA. Retrieved from <https://arohatgi.info/WebPlotDigitizer/>
- Rubec, C. D., & Hanson, A. R. (2009). Wetland mitigation and compensation: Canadian experience. *Wetlands Ecology and Management*, *17*, 3–14.
- Shafer, D. J., & Roberts, T. H. (2008). Long-term development of tidal mitigation wetlands in Florida. *Wetlands Ecology and Management*, *16*, 23–31.
- Simmons, B. A., Wilson, K. A., Marcos-Martinez, R., Bryan, B. A., Holland, O., & Law, E. A. (2018). Effectiveness of regulatory policy in curbing deforestation in a biodiversity hotspot. *Environmental Research Letters*, *13*, 124003.
- Sinclair, S. P. (2018). *The role of social factors in complex decision-making processes* (PhD Thesis, Imperial College London). Retrieved from https://www.iccs.org.uk/sites/www.iccs.org.uk/files/inline-files/Sinclair-S-2018-PHD-thesis_0.pdf
- Sonter, L., Tomsett, N., Wu, D., & Maron, M. (2017). Biodiversity offsetting in dynamic landscapes: Influence of regulatory context and counterfactual assumptions on achievement of no net loss. *Biological Conservation*, *206*, 314–319.
- Sonter, L. J., Barnes, M., Matthews, J. W., & Maron, M. (2019). Quantifying habitat losses and gains made by US Species Conservation Banks to improve compensation policies and avoid perverse outcomes. *Conservation Letters*, *12*, e12629.
- Steffen, W., Broadgate, W., Deutsch, L., Gaffney, O., & Ludwig, C. (2015). The trajectory of the Anthropocene: The great acceleration. *The Anthropocene Review*, *2*, 81–98.
- Stone, E. L., Jones, G., & Harris, S. (2013). Mitigating the effect of development on bats in England with derogation licensing. *Conservation Biology*, *27*, 1324–1334.
- Sudol, M. F., & Ambrose, R. F. (2002). The US Clean Water Act and habitat replacement: Evaluation of mitigation sites in Orange County, California, USA. *Environmental Management*, *30*, 0727–0734.

- R Core Team (2018). *R: A language and environment for statistical computing*. Vienna, Austria: R Foundation for Statistical Computing.
- Teels, B. M., Mazanti, L. E., & Rewa, C. A. (2004). Using an IBI to assess effectiveness of mitigation measures to replace loss of a wetland-stream ecosystem. *Wetlands*, 24, 375.
- Theis, S., Ruppert, J. L., Roberts, K. N., Minns, C. K., Koops, M., & Poesch, M. S. (2019). Compliance with and ecosystem function of biodiversity offsets in North American and European freshwaters. *Conservation Biology*. <https://doi.org/10.1111/cobi.13343>
- Thorn, S., Hobbs, R. J., & Valentine, L. E. (2018). Effectiveness of biodiversity offsets: An assessment of a controversial offset in Perth, Western Australia. *Biological Conservation*, 228, 291–300.
- Van den Bosch, K., & Matthews, J. W. (2017). An assessment of long-term compliance with performance standards in compensatory mitigation wetlands. *Environmental Management*, 59, 546–556.
- Walker, S., Brower, A. L., Stephens, R., & Lee, W. G. (2009). Why bartering biodiversity fails. *Conservation Letters*, 2, 149–157.

SUPPORTING INFORMATION

Additional supporting information may be found online in the Supporting Information section at the end of the article.

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