Tourism in Sub-Global Assessments of Ecosystem Services

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Abstract

Published in 2005, the United Nations Millennium Ecosystem Assessment (MA) stressed that influencing governments, businesses and communities to address the supra-national challenge of limiting biodiversity loss and ecosystem degradation requires a fuller understanding of the range of values and benefits people derive from ecosystems, including tourism. The MA was informed by, and has shaped, several conceptually- and methodologically-distinctive sub-global assessments (SGAs) of ecosystem services. Through content analysis, this paper is the first detailed examination of how tourism features in 14 extant SGAs identified in a database held by a major supra-national environmental organization. Although the SGAs should have incorporated the widest range of specialist subject expertise, expert tourism scholars played only peripheral roles in producing them even for territories where tourism is a significant land use. The SGAs examined did not benefit from the extensive body of knowledge relating to sustainable tourism. Limited portrayals of tourism restrict the capacity of SGAs in their current format as management solutions. It is also contradictory to the ethos, principles and purpose of ecosystem assessments. With the ecosystem services perspective set to become more important to policy and decision making, the paper argues for greater incorporation of recent progress in sustainable tourism in ecosystem assessment.

Key words

Ecosystem services, sub-global assessment, sustainable tourism.
Introduction

Global economic development has been accompanied by natural resource depletion, placing ecosystems around the world under intense pressure (European Commission, EC, 2014). The United Nations Millennium Ecosystem Assessment (MA) published in 2005, was a clarion call for those seeking to influence governments, businesses and communities to address the supra-national challenge of sustainable development especially through limiting biodiversity loss and ecosystem degradation. The MA has been informed by, and has subsequently shaped, a series of conceptually- and methodologically-distinctive sub-global assessments (SGAs) at various spatial scales. The focus of these assessments has been on conveying the dynamic and changing contributions that ecosystems make to human well-being. They do this by charting states and trends in the provision of ‘ecosystem services’ and the associated (predominantly economic) values and benefits arising from these (Ash, Blanco, Brown, Garcia et al., 2010; Chan, Satterfield and Goldstein, 2012). Among SGAs tourism is routinely considered where it is an appropriate major ‘local’ land use. Assessment of ecosystem services at the global and sub-global level has assumed further importance with the establishment in 2012 of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services (IPBES). Analogous to the Intergovernmental Panel on Climate Change (IPCC), IPBES will conduct ecosystem service assessments at the global scale and in four continental regions (Diaz, Demissew, Joly, Lonsdale and Larigaudiere 2015). However, to date there has been no systematic analysis of how tourism features in SGAs, a lacuna which this paper sets out to address. This is a notable oversight. Tourism is a complex form of human activity that defies simple definition or straightforward categorisation (Hall, 2005). As tourism scholars have long recognised (Mathieson and Wall, 1982), the multiple roles and outcomes of tourism in economy, environment, society and culture require careful problematization and reconciliation,
not least in approaches to sustainable tourism (Mowforth and Munt 1998; Duffy 2015; Hunt, Durham, Driscoll and Honey, 2015).

The aim of this paper is to present the first critical examination of how tourism features in the assessment of ecosystem services. It does this through a content analysis of the published record of SGAs. Through their integration of science, policy and practice on the themes of economics, conservation and development, SGAs should be created and compiled on the basis of inter- or trans-disciplinary expertise (Braat and de Groot, 2012; Guimaraes, Balle-Beganton, Bailly, Newton et al. 2013). The principle is that they should able to ‘bridge gaps between different academic disciplines and research communities’ (Milcu, Hanspach, Abson and Fischer, 2013, p.44). A specific objective of this paper is to examine the extent to which SGAs in the many territories where tourism is a major land use, benefit from the extensive, well-established, mature body of knowledge relating to sustainable tourism (Ruhanen, Weiler, Moyle and McLennan, 2015). After all, ecosystem services thinking and sustainable tourism share the same underlying logic and goals for sustainable development as well as intellectual and philosophical genealogies stemming from the Brundtland Commission (Saarinen and Rogerson 2014; Duffy 2015). Furthermore, constructs from ecosystem services have started to appear in management solutions for sustainable tourism (Simmons, 2013; Whitelaw, King and Tolkach, 2014; Mayer, 2014). However, this paper will argue that the tourism academy –in the sense of an epistemic community with well-established professional norms, concepts and standards- has been peripheral to the production of extant SGAs. Before these arguments are elaborated, we first review recent academic progress on ecosystem services by its particular epistemic community. In particular, we focus on where tourism may be conceptually positioned by scholars adopting this approach. This discussion frames the content analysis later in the paper in so far as the coding is in part based on key features to have emerged from the literature.
**Literature Review: Ecosystem Services, Culture and Tourism**

Ecosystem services is a concept designed in the practice of inter-disciplinary resource management. Its origins are in natural- and social-scientific debate concerning nature conservation, biodiversity protection and environmental sustainability, in particular at the interface between ecology and economics (Constanza, d’Arge and de Groot, 1997; Constanza, de Groot, Sutton et al., 2014). The core purpose of the concepts is to convey the many and varied contributions that ecosystems can make to human well-being. As such, the concept belongs within a utilitarian tradition of research and policy development (Fish, 2011, Fish, Church and Winter 2016). The natural world assumes significance for decision-making on the basis of what people use and derive from nature. In practical terms the main focus of an ecosystem services perspective is to characterise and build evidence for this link between ecological phenomena and human quality of life. An epistemic community of researchers and practitioners has emerged developing increasingly standardized ways of identifying, modelling and valuing these services to inform decision-making at a variety of spatial scales (Ash et al., 2010).

Clearly then, one important aspect is to convey human dependency on ecosystems without being purely ‘resourcist’ in manner. By promulgating the language of ‘services’ and ‘benefits’ there is a deliberate shift away from viewing nature as merely an object to be protected against exploitative and/or disruptive human behaviour, towards harnessing its life-giving and life-enriching qualities (MA, 2005). The natural environment is not only generative of fundamental benefits to human welfare—such as providing sustenance and securing livelihoods— but also more interpretative aspects of social and cultural re-production (Bhagwat, 2009). For instance, these may range from conveying the historic and religious significance of ecosystems to the building of social capital within communities. Reflecting this, an increasingly harmonised threefold framework for categorising ecosystem services has evolved, namely:
• **Provisioning services**: the products obtained from ecosystems, including food, fibre, fuel, genetic resources and fresh water;

• **Regulating services**: the maintaining functions of ecosystem processes, including regulation of air quality, climate, water quality and natural hazards;

• **Cultural services**: the contributions ecosystems make to processes of life enrichment, such as cultural identity, cognitive development and aesthetic experience; and

Comprehensive reviews have charted the development of this distinctive perspective (e.g. Potschin and Haines-Young, 2011; Milcu et al., 2013) which has also attracted lively critical debate (Robertson, 2012; Schröter et al., 2014). Of particular relevance here is that such thinking now cuts across, and increasingly harmonizes, diverse aspects of environmental planning (Burkhard, Petrosillo and Constanza, 2010, Bryce et al. 2016) centred on two arenas of science-policy innovation. First, ecosystem services are the operational concept informing the practice of ’ecosystem assessment’ (Ash et al., 2010); that is, scientists and policy-makers gauging states and trends in ecosystem service provision in order to elevate priorities for the natural environment across government, business and civil society. This includes the influential pan-global Millennium Ecosystem Assessment (MA, 2005), commissioned by the United Nations Secretary-General and the series of related SGAs which provide the subject matter for this paper.

Second, ecosystem services underpin the so-called ‘ecosystems approach’ to natural resource management. Formatively associated with the International Convention on Biological Diversity (CBD, 1993) which adopted it as its primary framework of action, the ecosystems approach is essentially a set of principles for embedding consideration of ecosystem services into decision-making. This includes encouraging inclusivity and cross-sectorality, promoting adaptive management and
local solutions as well as ensuring that the market and non-market value of ecosystem services are fully accounted for in policy, plan and project design (Ash et al., 2010).

Taken together, ‘ecosystem assessment’ and an ‘ecosystem approach’ are mutually-reinforcing contexts for the promotion of ecosystem services in the mediation of environmental futures: the former rooted in broad, science-informed advocacy of the natural environment; the latter in the methods and mind-sets of delivery. Both ideas actively cross-fertilize each other, sometimes to the extent that they are viewed as one and the same thing. This is not least because applications of the ecosystem approach often demands the process of ecosystem assessment. Key components of assessment methodologies (which structure the analysis below) include measuring the condition and trends of ecosystem service provision; identifying the drivers of change and their impact variables to show cause-and-effect relationships between services and human well-being over time (including through the scenario heuristic); and importantly, providing commentary on the multi-dimensional responses across state, business and civil society required to affect change in normative terms (MA, 2005; Ash et al., 2010).

For an approach that ranges over, and seeks to incorporate, diverse connections between ecosystems and well-being, there are strong grounds for incorporating tourism beyond the general need to incorporate diverse disciplinary and sectoral interests (Braat and de Groot, 2012). In principle, tourism can be conceptualised in three main ways from an ecosystem services starting point. First, tourism as a sector capitalises strongly on the values people have for natural environment assets. These may be the individual economic valuations people pay for the natural environment they access for specific tourism activities or the shared dimensions of value which arise from the cultural, social or group experiences gained through tourism. This concern with values has also been a long standing concern of the sustainable tourism literature that has stressed how the values of different communities affect how they perceive the benefits of tourism for sustainable development (Mowforth and Munt 1998; Duffy 2015). The
way in which ecosystems are managed therefore has a fundamental bearing on the tourism sector’s capacity to reproduce itself as a particular set of mediations and experiences of nature. In this sense, tourism businesses may be viewed as economic beneficiaries of the contributions that ecosystem services make to well-being. This was the focus of the MA and it relates to the changing benefits humans gain from interactions with nature including certain forms of tourism experiences such as ecotourism (MA 2005). Second, tourism is predicated on resource dependencies across the full range of ecosystem services. Thus, it may be viewed as a driver of change on ecosystems service production in positive and negative ways simultaneously. For instance, mass tourism developments can fundamentally degrade terrestrial and marine ecosystems and their ability to support not just cultural services but also provisioning and regulating services as well. Finally, tourism is a context in or through which (other) services from ecosystems are actively made. It is a sector that creatively constructs and re-constructs the well-being roles ecosystems play in peoples’ lives. Hence, tourism is more than simply a conduit of benefits from, and impacts upon ecosystems; it is a process in which ecosystem services are produced by way of particular tourist products and ‘offers’. For example, farm tourism now functions as a part of pluri-active land management system which shape the way food as a provisioning service is produced (Busby and Rendle 2000).

Notwithstanding distinctive vocabularies and different intellectual terms of reference, these core ideas are hardly novel for tourism scholars and the body of knowledge they have produced (that is, the ‘tourism literature’). There is a strong conceptual fit between the idea that ecosystem services are mediated, consumed and created through tourism with existing tourism studies, for instance, on the multiple relationships between tourism, health and well-being in both physical and psychological ways (cf. Nyaupane and Soudel, 2011; Chen and Petrick, 2013; Ram, Nawijn and Peeters, 2013; Bryce et al. 2016). Furthermore, the concern of sustainable tourism scholars with
sustainable development more generally (Hall and Page 2014) also clearly resonates with the concern of ecosystem assessment. IPBES was specifically established as a global initiative to enhance the science policy interface between sustainable development, human well-being, ecosystem services and biodiversity (Diaz et al. 2015). Still, understanding of tourism starting from an ecosystem services perspective is currently heuristic and generally tentative. Several analyses of the peer-reviewed literature locate recreation and/or tourism as central to consideration of cultural ecosystem services (CES) (Hernandez-Morcillo, Pleininger and Bieling, 2013; Geijzendorffer, Martin-Lopez and Roche 2015; Bryce et al. 2016; Fish, et al. 2016). For Fish (2011, p.674), the focus tends be on ‘a rather underwhelming and predictable set of activities, such as types and patterns of recreation and (undertheorized) appeals to aesthetic value’. Part of the problem is that they encompass processes apparently less discernible, and therefore less studied, than those belonging in other classes of ecosystem service (Pleininger, Dijks, Oteros-Rozas and Bieling, 2013). The result is that tourism in general -and sustainable tourism more especially- as they are understood by tourism scholars occupy places at the margins of this discourse.

Research problem, methods and data sources

Despite such issues, in principle tourism deserves attention in SGAs that are inspired by, and follow, the MA (2005). Based on the discussion so far, this paper now examines how tourism has featured in SGAs produced by a range of governments and non-governmental organisations by addressing three fundamental questions.

First, how is tourism understood in SGAs? Not only does this pertain to how tourism is conceptualized but also which particular forms of, or approaches to, tourism are identified. For instance, given their shared genealogies it may be reasonable to expect discussion of various facets of sustainable tourism to feature prominently in
SGAs. Second, the ecosystem services framework is concerned with defining and measuring -often using indicators- the benefits to well-being that flow from using and interacting with particular environments as well as identifying policy responses that can govern and help sustain these. This raises the question of what type(s) of well-being benefit are commonly identified with tourism and, more specifically, how is tourism measured and valorised? Within this context, we may reasonably expect the significant progress in developing sustainable tourism indicator sets to have been embraced by SGAs (Miller and Twining-Ward, 2005; Torres-Delgado and Saarinen 2014). Finally, from an epistemological perspective, how has knowledge about tourism been produced in SGAs? The body of knowledge about tourism has grown rapidly in the past two decades as has the size and membership of the tourism academy which spans the arts, humanities and social sciences (Coles, Hall and Duval 2016). Hence, how far have published SGAs drawn on this published record and the scholarly networks that surround it?

An extensive content analysis was conducted in 2012 and 2013 which focused on CES within SGAs, that is, where tourism is routinely considered (Hernandez-Morcillo et al. 2013; Bryce et al. 2016). Although in principle tourism may function in, or contribute to, provisioning or regulating services in practice none of the SGAs interpreted it in this way. The listings of a major United Nations Environment Programme (UNEP) database were reviewed. At the time of writing, this contained entries for 82 ecosystem service assessments. The MA was not only a pan-global assessment but also multi-scalar in design. It involved a series of 18 UN-approved and 15 UN-Associated SGAs at different geographical scales, including pan-national regions, nation states and urban regions (MA, 2005), each of which was inscribed in this UNEP database. With some ‘local’ adaptation, both types employed the same conceptual approach and empirical apparatus. By 2012 UNEP had also gathered information of varying scope on 47 further SGAs postdating the MA process, with two far more
extensive national-scale assessments for Japan (Duraiappah et. al 2012) and the UK (UK NEA, 2011).

Ultimately, 28 SGAs were retained for detailed content analysis (Table 1). Fourteen of the 82 were discarded, since UNEP had insufficient information to sustain detailed analysis. Its database holdings were in summary form or secondarily derived from other sources, such as related policy documents or a UNEP survey of assessment authors. Eleven assessments were eliminated because they entirely overlooked CES and therefore tourism. For a further 29 discounted assessments, it was impossible to obtain sufficient (English-language) material to undertake rigorous analysis. Several SGAs in the database, including UN-Associated and -Approved assessments, were no longer publically-available on web sites or in paper form. Some assessments were excluded because they were not in English or they were available in other languages in which the researchers were not competent (e.g. Vietnamese and Lithuanian).

The 28 remaining SGAs were read, coded manually and corroborated by two independent investigators. Of the 28 SGAs only half mentioned tourism as a CES. The other 14 SGAs were nevertheless inspected to ensure that substantial material on tourism and ecosystem services was not overlooked in error. Eleven mentioned tourism in passing when considering the processes that caused alterations in ecosystem services. For example, in a 142-page Canadian study of ecosystem status and trends, tourism is cited just once as a factor contributing to coastal erosion (FGPTC 2010).
There are numerous approaches to content analysis (Krippendorf, 2013). This research combined a series of basic quantitative measures with a qualitative approach to the textual content of the SGAs (Schreier, 2012). Standard indicative metrics were compiled (i.e. word counts, page length, positioning of information, authorship and so on) as surrogate indices of issue importance (Table 3). Separate in-depth readings were made by two researchers who then compared their findings. Three major themes emerged from the readings: conceptualisations of tourism in SGAs; tourism in measures in SGAs; and tourism as drivers of change that affect ecosystems and related services. These should not be surprising. After all, eleven of the 14 SGAs were UN-approved or -associated. Although the reading process started without pre-ordained views, these themes reflect emphases in the protocols for undertaking ecosystem assessments as recommended by the MA (2005) and elaborated in the related literature. These include: utilising a conceptual framework; measuring ecosystem services; and identifying the implications for human well-being of changes in ecosystem services and how these are affected by key drivers of change (Ash et al., 2010). The three themes are discussed in turn below after an assessment of how tourism was present in SGA texts more generally.

**Analysis: tourism in SGAs**

As Tables 2 and 3 indicate, there were marked differences in the size and scope of the 14 SGAs that were analysed in depth because they mentioned tourism. One major national assessment, the UK report at 1,452 pages, is much longer than any other document and hence measures are also provided for its shorter synthesis report. This is 79 pages and can be compared to most of the other assessments which range in length from 21 to 166 pages (Table 3). Nine documents comprised final reports among which three were given slightly different titles (cf. Coastal British Columbia, Southern Africa and Portugal). Among the other five assessments, three were summary reports as they were the only
documents available; one (Glömma Basin, Norway) was a pilot assessment for a proposed national study; and another (Switzerland) was termed an assessment framework and methodology but included analysis and data on ecosystem services.

The number of words devoted to topics can be a relatively limited indicator of issue importance and here this was affected by the varying length, layout and production values of each SGA. 'Page mentions' proved more useful and in eight of the 14 texts, 9% or more of the pages focus on CES. Some assessments, such as the Caribbean Sea and San Pedro de Atacama (Chile), had entire sections on tourism. The large-scale UK assessment contains a list of 124 key findings of which 11 are focussed on tourism. Nevertheless, in general within each SGA except Coastal British Columbia (Canada) tourism is considered both within and beyond the section on CES. For example, tourism is mentioned on 69% of pages in the Greater Jakarta Bay (Indonesia) assessment whereas the section on CES constitutes only 9% of that text. Consideration of tourism beyond CES typically occurs in sections on drivers of change (in which tourism is considered as a key factor that has recently altered other ecosystem services) and in scenario analyses which outline future decision options. For example, the Caribbean Sea assessment contains a 29-page annex on future scenarios of which 20 pages make mention of tourism as an ecosystem service per se. In fact, in eight of the 14 assessments, tourism is mentioned on 23% or more of the pages. Some care must be taken not to over-read the importance of tourism. With the notable exception of the UK assessment, the terms ‘tourism’, ‘recreation’ and ‘leisure’ are used in a conceptually-conflated, undifferentiated manner. Other assessments use the term ‘recreation’ in their accounts of CES and (day) visitors are often confused with, or described conveniently as, ‘tourists’.

A marked diversity is evident in the spatial extent, habitats, degree of urbanisation and development of tourism in the locations covered by the SGAs. They include locations in all continents except Antarctica, a range of states in both the global
north and south, and contrasting tourism destinations. These vary from San Pedro de Atacama (Chile) that receives 50,000 visitors per annum to a Caribbean Sea area that attracts 25 million visitors a year. In some locations, such as the Caribbean Sea, tourism is a central element of the local economy. In others, such as the Laguna Lake Basin (Philippines), tourism development is limited while in the Glömma Basin (Norway) it is a relatively small but growing industry. Significantly, tourism is acknowledged in ecosystem assessments even where it is not a major sector in the economy. However, more space in assessments is devoted to tourism where it has a major presence within multiple destinations (Table 3). For instance, tourism is one of only two ecosystem services considered in-depth in the Caribbean study (the other is fisheries). The Northern Range of Trinidad assessment describes tourism as part of CES for which it reports demand is already high and expanding. By contrast, CES comprises less than 3% of page length for Coastal British Columbia (Canada) or the Coffee growing regions of Colombia, and hence tourism receives limited consideration. With such variations in attention, it is not surprising that in some studies tourism is discussed in a very generalised manner, a point to which we return below.

Each assessment was undertaken between 2002 and 2012 and some indicate they have learned from their predecessors (e.g. UK in 2011). There was no evidence that the date of assessment affected the scope or depth of coverage of tourism. There are both early (e.g. Glömma Basin, Norway in 2002) and later (e.g. Japan 2012) assessments containing significant volumes of material on CES and tourism. Despite its apparent importance, there was little evidence that recognised tourism experts (i.e. from within the established tourism academy) contributed to the compilation of assessments. Guidance for undertaking ecosystem assessments proposes that expert authors should be drawn from a wide variety of disciplinary backgrounds and should include practitioners as well as academics (Ash et al., 2010). As a result, authorial teams are often large in number and they include lead authors as well as extended teams of co-
authors for specific chapters or sections of the assessment. In some cases, these arrangements are complex. For example, the Japanese assessment cites 170 authors, many working on the six cluster reports for specific geographical sub-areas of Japan, and of these 39 contributed to the main national assessment.

For just eight of the 14 assessments, sufficient detail of the author teams was provided. This means that for over a third of the assessments in which tourism is invoked as a significant activity, it is impossible to ascertain who is responsible for drawing conclusions about tourism and on the basis of what credentials. Indeed, only four of the eight SGAs providing author details identified tourism and/or recreation specialists in their teams. Usually these were a small number of specialists within large teams so that, of the 228 authors whose backgrounds were examined, just ten (4.4%) authors claimed tourism and/or recreation expertise. Of these five were associated with a university and the other five were from practitioner/consultancy backgrounds. Tourism analysis was routinely undertaken by authors whose disciplinary backgrounds were described primarily as being in economics or ecology. The involvement of economists in part reflects the (monetary) valuation of services as a key component of ecosystem assessments especially in some SGAs (Ash et al., 2102).

**Conceptualisations of tourism in ecosystem assessments**

The broad analysis of content reveals clear variations in the level of interest in tourism. In part, this is due to the flexibility imbued in the UN’s guidelines for undertaking ecosystem assessments to acknowledge location-specific contexts. In principle, though, this should have resulted in the identification of multiple and varied tourism types, as they relate to market characteristics in the areas under assessment.

Some assessments, such as that for Glömma Basin (Norway) discussed multiple types of tourism in different habitats. However, they were in the minority and, instead as Table 3 indicates, the conceptual approach led to a narrow focus on specific types of
tourism. As may perhaps have been anticipated, there was a heavy emphasis on ‘ecotourism’, broadly conceptualised, which is considered in eight of the 14 assessments. The significance of animal species is manifest in the fact that six assessments discussed hunting as a form of tourism reliant on ecosystems. Just three assessments considered farm based- or agri-tourism as a significant feature of pluri-active agricultural systems and economies. In some assessments, such as that for Southern Africa, eco-tourism is considered alongside nature-based tourism. The differences between the two are not explicated, although nature-based tourism is often inferred to involve hunting and fishing.

No assessment presented a clear rationale for focusing on ecotourism. Rather, its relevance is assumed: since the conceptual framework seeks to highlight the benefits humans gain from ecosystems, ‘ecotourism’ must be the most appropriate moniker. This is without any further detailed scrutiny of the term and its meaning, nor cross-referral to long-standing definitional discourse among the tourism academy (Fennell, 2001; Donohoe and Needham, 2006). Discussions of ecotourism in SGAs are often relatively optimistic about its economic and environmental roles compared to more critical academic analyses (cf. Butcher, 2005; Weaver and Lawton, 2007; Fennell, 2008; Duffy 2015; Hunt et al., 2015). Assessments overlook that in some locations the commodification of nature for ‘ecotourism’ can result in negative, as well as positive, outcomes for local communities and ecosystems (Stronza and Gordillo, 2008; Hunt et al., 2015).

Emblematic of these issues, ecotourism is devoted 6% of the word length of the India Urban Resource assessment where simplistically it is argued to have certain advantages over other forms since tourists are willing to pay entry fees to such sites which provide incomes to local communities and small businesses. The report concludes that,
‘eco-tourism inside the city would be a big business worth [Rupees] Rs. 1 million at Rs. 10 per person per morning trail. Its economic contribution in terms of health security & cultural satisfaction would be recognised as much higher.’

[Naturalist, 2005, p.16]

Similarly, the Southern Africa assessment argues that nature-based tourism can contribute to management that has positive implications for conservation and improvements in a range of ecosystem services concluding that the,

‘combination of private game farming, trophy hunting and nature-based tourism has expanded the conservation estate to a remarkable degree’.

[Biggs, Bohensky, Desanker et al., 2004, p.29]

Within tourism studies, however, there has been extensive problematization of ecotourism and nature-based tourism (Weaver and Lawton, 2007; Fennell, 2008; Newsome, Moore and Dowling, 2012; Duffy 2015). Ecotourists and nature-based tourists do not form homogenous or singular groups as they are generally portrayed in SGAs. Rather, tourism research has identified different sub-types of ecotourists and nature-based tourists (Mehmetoglu, 2007; Arnegger, Woltering and Job, 2010; Deng and Li 2015). Not only do these types vary in their connections to nature but also the benefits they gain from, and contribute to, ecosystems. As a result they pose quite distinctive environmental management challenges that are not acknowledged in SGAs.

Finally, beyond ‘ecotourism’ other interpretations of tourism are coarse and undifferentiated. For instance, despite their emphasis on CES, some SGAs entirely overlook other major tourism market segments that are clearly dependent on provisioning services and local ‘natures’, such as food and drink-related tourism (Hall and Gössling 2013) or agri-tourism (Sznajder, Przezborska and Scrimgeour 2009;
Cruise tourists are subject to the same reductionist tendencies as ecotourists despite similar progress in tourism studies in identifying different types and their implications for destinations (Papathanassis and Beckmann, 2011). By and large analyses of how tourism is related to different forms of material and non-material human well-being (cf. Chen and Petrick 2013; Filep 2014; Romagosa et al., 2015) are simply overlooked and not invoked in SGAs. Indeed, discussion of well-being as it relates to tourism is one-dimensional in the few instances where it appears.

The Caribbean Sea assessment focuses on the economic aspects of well-being; the San Pedro de Atacama (Chile) study argues that the major impact of tourism on well-being is linked to changes in traditional ways of life; and the Northern Range of Trinidad assessment takes a far more extensive view but still just identifies how tourism contributes to well-being through economic activity, recreation and social relations.

**Tourism in measures in SGAs**

Some standard measures of tourism activity are integrated in SGAs but they are used unquestioningly manner without consideration of the critical issues or limitations identified in tourism research.

In the nine SGAs using basic measures, the most common are visitor numbers, expenditure and tourism employment. For example, the San Pedro de Atacama (SPA) assessment confidently asserts that,

'the SPA is one of Chile’s best known tourist destinations.....amongst the top eight destinations in the country for foreign tourists. Over fifty thousand people are estimated to visit the municipality every year and over 60% of these are foreign tourists (mainly Europeans, particularly French tourists). In 2000, it was estimated that tourism in SPA brought in over five million US dollars in foreign
currency. Due to its importance, part of the SPA municipality was declared a Zone of National Tourist Interest.’

[RIDES, 2005, p.21]

In some assessments, more complex economic valuations are provided in order to highlight the economic benefits of improved environmental management and conservation. For example, the UK assessment incorporates detailed estimates of non-market forms of tourism and recreation (e.g. day visits to the countryside) alongside the analysis of market goods based on ecosystems (e.g. food and timber). This makes a significant difference to the resulting economic values and policy options in terms of measures to facilitate land use change in the UK.

Other SGAs lack a similar level of analysis and instead their reliance on basic measures has connotations for how tourism is understood and factored into decision-making. This is because, with the exception of the Laguna Lake Basin (Philippines) assessment, such measures all show apparent increases in tourism: in other words, the condition of CES is improving. However, the latter (visitor numbers) may point to an altogether different story (about pressures on ecosystems and carrying capacity). Indeed, few assessments simultaneously demonstrate how tourism interactions with ecosystems are more complex, generating both benefits and threats. One exception, the Greater Jakarta Bay (Indonesia) assessment notes that,

‘Tourism activities could economically bring benefits not only to the local government but also to local communities. However, the number of tourists may need to be controlled as the increase in number of diving and other tourist related activities on the islands (including coastal development to support tourism) can have a negative impact on the condition of ecosystems surrounding tourist areas.’
Overall, consideration of the negative dimensions of tourism in assessments is restricted to instances where expansion has been relatively rapid, such as cruise tourism in the Caribbean Sea or coastal tourism in the Portugal assessment. Furthermore, although each assessment views tourism as a benefit humans gain from ecosystems, only eight assessments acknowledge the potential or actual contribution of tourism to environmental degradation.

These discussions of the contrasting impacts of tourism raise another challenge that is routinely missing from assessments and which reflects wider problems with CES (Chan, Satterfield and Goldstein, 2012; Bieling and Plieninger, 2013; Hernandez-Morcillo et al., 2013, Fish, et al. 2016): that is, what constitutes an appropriate measure(s) of tourism and recreation as ecosystem services (since growth in visitor numbers may create significant dis-benefits)? Seven assessments use standard, readily-available tourism visitor and employment data usually gained from national or local sources. For example, the measures for the Glömma Basin (Norway) assessment of 90,000 coastal fishing tourists and 350,000 pleasure boats located on the coast come from national statistics. The Greater Jakarta Bay (Indonesia) survey draws general conclusions about tourism and recreation based on a local survey of the number of visitors and holiday cottages on the islands in the bay. Just two of the assessments make use of recognised international sources of tourism data to evidence their positions. For example, the Caribbean Sea assessment employs data from the World Travel and Tourism Council, the Caribbean Tourism Organisation and Oxford Economic Forecasting.

Reliance on national and local sources is problematic. As is well established in tourism studies, many secondary data sources, such as those produced by national tourism organisations have significant limitations (Page and Connell, 2009; Lam and
McKercher, 2013; De Cantis, Parrocco, Ferrante and Vaccina, 2015) while the design and interpretation of local-level studies of the economic impacts of tourism is highly contestable (Crompton, 2006). Moreover, different visitor types and segments can generate very different interactions with local landscapes and environments (Dolnicar, 2010; Arnegger et al., 2010). Simple measures of tourism activity of the type used in SGAs almost inevitably gloss-over the findings of much tourism research which reveals how complex arrays of economic activities combine to produce ‘visitor experiences’ that tourists consume, and all of which should be addressed in considering management options (Hall and Page, 2014; Weidenfeld, Butler and Williams, 2011, 2014; Benur and Bramwell 2015).

**Tourism as driver of change**

Many assessments discuss the challenges tourism raises for developing management options in ecosystems and habitats. For example, the Northern Range of Trinidad assessment notes how significant conflicts arise from the management of natural environments to deliver different ecosystem services including mass tourism arguing that,

‘High and growing demand for the use of the coastal resources for recreation, tourism enterprises, fisheries, anchorage and sea transportation, and conflicts among these activities for use of limited resources, have implications for sustainable management.’

[Northern Range Assessment, 2005, p.xvi]

Such a comment also exposes a key conceptual challenge of addressing tourism within the ecosystem services framework. In eight of the 14 assessments, tourism in general was identified as both a service producing benefits for humans and a driver or cause of
‘ecosystem change’. For example, in the Japanese assessment tourism is identified as a significant impetus for a major transformation,

‘There is a renewed and growing interest within Japan to revive these traditional rural landscapes called satoyama and satoumi but within the context of modern Japan.....the increasing demand for eco-tourism has been a driving factor for satoyama and satoumi renaissance.’

[Duraiappa, Nakamura, Takeuchi, Watanabe and Nishi, 2012, p.2]

Nine of the 14 assessments discuss tourism using the term ‘driver’ or ‘driving force’. This is because such terms are part of the MA apparatus which suggests SGAs should identify drivers of change to inform policy debates. Drivers are natural or human-induced factors that directly or indirectly cause a change in an ecosystem service. Conceptually, they may be divided into primary/indirect drivers (e.g. population change) and proximate/direct drivers which are pressures (e.g. harvesting). They may be further differentiated into drivers that are endogenous or exogenous to the territory of the SGA. The nine SGAs that consider tourism as a driver of change adopt varied approaches. Two assessments (Caribbean Sea and Greater Jakarta Bay) consider tourism within socio-political drivers since it is the lack of regulation of tourism that is perceived to be degrading natural environments. In the other seven assessments mainly economic processes are discussed. Changes in demand or market segments lead to a growth in tourism which drives change in a range of ecosystem services. For example, the SGA for Portugal identifies international tourism markets as an exogenous driver of change, while demand for second homes and the activities of the construction sector are an endogenous driver degrading certain coastal ecosystems. Similar to the discussion of ecotourism and data sources above, as a driver of change tourism is treated in a relatively simplistic and unproblematic manner when compared to the tourism
literature. For instance, while different forms of demand that contribute to tourism as a
driver are routinely ignored, where assessments (e.g. Portugal and Caribbean Sea)
compare demand, they do so coarsely only in terms of domestic and international
tourism. Indeed, the wide array of ‘drivers’ of change in tourism demand, not to say the
methods and approaches to forecasting tourism demand that are much debated in
tourism studies (Song and Li, 2008; Athanasopoulos, Hyndman, Song and Wu 2011),
receive very limited attention within extant SGAs.

**Discussion and Conclusion**

Since the publication of the MA, interest in ecosystem services has gathered momentum
globally. Several past SGAs of varying spatial scope have both informed, and been
shaped by, the MA. In turn, extant SGAs have inspired, and will act as a further
foundation for, subsequent SGAs. This paper has concluded that there is a lack of
engagement of SGAs with the concepts and arguments in the sustainable tourism
literature and also with the tourism academy more generally. This conclusion, however,
is not intended to be a complaint about the exclusion of tourism researchers from
ecosystem assessments. Instead, the paper has highlighted that there are synergies
between the ecosystem service perspective and sustainable tourism research, and that
clear opportunities exist within the conceptual framework of ecosystem assessment to
incorporate tourism and tourism-related activities where they represent appropriate
benefits to human well-being linked to ecosystems and biodiversity. There is also clear
scope for SGAs to be informed by recent progress in sustainable tourism. Furthermore,
recent developments at the global level through IPBES present opportunities both for
the tourism academy and tourism organisations more generally. The IPBES initiative is
supported by 124 governments and the IPBES global and four regional assessments
have a specific goal to involve wide ranging expertise in the assessments from different
disciplines in social science as well as natural science (Larigauderie, Stenseke and Watson 2016). Tourism researchers should see this as an opportunity to become more closely involved in these assessments. In addition, IPBES is required to ensure its utility through extensive stakeholder engagement and capacity building which are part of the assessment process (Diaz, et al. 2015). Tourism governance organisations and tourism NGOs should take advantage of these processes as a route to influencing how tourism is treated in the major IPBES ecosystem assessments and any related policy outcomes.

The SGAs considered in this paper, however, suggest that to date there are only limited signs that tourism is being adequately addressed in this way of thinking. In half of the 28 SGAs available for inspection here tourism was incorporated in the assessment of CES; that is, where practical guides (Ash et al., 2010) and academic commentaries (Plieninger, et al., 2013) argue it is most appropriate and hence likely to appear. Among these existing SGA texts tourism was understood and portrayed in a rudimentary manner through simplistic analysis and language. Routinely, tourism, leisure and recreation were conflated. At a basic level, there was little attempt to differentiate between tourists and (day) visitors either as beneficiaries from ecosystem services, or as potential drivers of ecosystem change. Some tourist types also reliant on provisioning ecosystem services and biodiversity like agri-tourists, food-and-drink tourists were overlooked altogether. Other categorisations of tourist types, such as eco-tourists and cruise tourists, appear used for convenience and neatness of labelling and in many cases can be questioned for their conceptual relevance and appropriateness. No reference was made to long-established debates and critical exchanges in sustainable tourism, nor to issues such as drivers for tourist behaviour and tourism demand.

The outcome is a series of limited representations of tourism in SGAs that are at best limiting, at worst potentially misleading when both the principles and the purposes of ecosystem assessment are considered. Tourists are mainly seen as the source of
monetary benefits from ecosystem services for local communities while the social and
cultural roles of tourism are obscured. There is a heavy reliance on secondary data
sources to deliver simple tourism-related metrics and indices, without reference either
to the relative merits of national and international tourism data sets nor the difficulties
in producing robust indicators sets in sustainable tourism. Instead, such sources are
used in unproblematic ways without a thoroughgoing acknowledgement of their
limitations, and they tend to portray uncritically the commodifying of ecosystems for
tourism as apparent benefits. Overall, tourism in SGAs is viewed from a highly
utilitarian and instrumental perspective which, paradoxically, the wider ecosystem
approach was originally intended to overcome.

Hence, this is another instance of knowledge production and dissemination
about tourism beyond the usual subject-specific channels (Wardle and Buckley, 2014).
It also presents compelling evidence of a lack of interaction and knowledge exchange
between two significant fields of study to their mutual detriment. Tourism scholars
have played too peripheral a role in an applied approach that is fast assuming
paradigmatic status in environmental management (Potschin and Haines-Young 2011,
p.575). Within the United Nations policy architecture on global environmental change,
biodiversity is as important as climate change (Loreau et al., 2006) as IPBES
demonstrates. The tourism academy has contributed to shaping the global agenda on
climate change through the IPCC (Amelung, Moreno and Scott, 2008). However, it has
not been at the forefront of debate about biodiversity as it relates to ecosystem services,
although the United Nations World Tourism Organization has made persistent efforts to
promote biodiversity as a theme (UNWTO, 2014). This is a missed opportunity because
more powerful approaches to valuation have notable potential to contribute to
unfolding tourism debates on so-called ‘green growth’ and the ‘green economy’ (Hall
2013; DeLacy, Jiang, Lipman and Vorster, 2014; Duffy 2015).
Viewed from the position of ecosystem services, one of the central tenets (and relative strengths) of the ecosystem services perspective is that it is capable of delivering enhanced environmental management with strong and discernible links to human well-being. This should be achieved by, as fully as possible, documenting and extending understanding of the range of benefits from biodiversity and ecosystems, and how these play out in an array of uses of land and marine systems. There are very few environmental contexts tourism has not penetrated to one degree or another. In order to deliver assessments to such a wide specification, inter-disciplinary teams are necessary that will benefit from the widest range of relevant expertise and knowledge from across the social and physical sciences, arts and humanities. Hence, in this context it is all the more surprising that expert tourism scholars and mature bodies of knowledge, especially relating to sustainable tourism, have at best been overlooked, at worst excluded despite the undeniable importance of this form of human activity. Clearly this situation highlights some key areas for further research including investigations with those responsible for SGAs as to why tourism was in the past considered as outlined in this paper. There is a need to shape pathways in which sustainable tourism issues will be addressed in the IPBES assessments in the future alongside other industry sectors such as forestry, fishing and agriculture. There is also a need to understand how the consideration of tourism in ecosystem assessment relates to provisioning and regulating services as well as cultural. Such research would play a role in avoiding a more sobering prospect which is that the current state of affairs may result in erroneous conclusions, recommendations, and policy decisions about future tourism activity. Unless there is greater acknowledgement and utilisation of recent progress in tourism studies, not least in the field of sustainable tourism, the raison d'etre of the ecosystem approach is disputable.
References


Braat, L. and de Groot, R. (2012). The ecosystem services agenda: bridging the worlds of natural science and economics, conservation and development, and public and private policy. *Ecosystem Services, 1*(1), 4-15


Northern Range Assessment (2005). *Report of an assessment of the Northern Range, Trinidad and Tobago: People and the Northern Range. State of the environment report 2004*. Port of Spain: Environmental Management Authority of Trinidad and Tobago.


Table 1: Stages of selection process

<table>
<thead>
<tr>
<th>Stage</th>
<th>Criterion</th>
<th>Number removed</th>
<th>Number remaining</th>
</tr>
</thead>
<tbody>
<tr>
<td>Start</td>
<td>Assessments cited in UNEP database</td>
<td>-</td>
<td>82</td>
</tr>
<tr>
<td>Stage 1</td>
<td>Removal of assessments with insufficient information to identify the ecosystem services analysed</td>
<td>14</td>
<td>68</td>
</tr>
<tr>
<td>Stage 2</td>
<td>Removal of assessments that did not consider CES</td>
<td>11</td>
<td>57</td>
</tr>
<tr>
<td>Stage 3</td>
<td>Removal of assessments for which access in the English language was unavailable or English-language material was a very short summary of less than 2 pages.</td>
<td>29</td>
<td>28</td>
</tr>
<tr>
<td>Sample</td>
<td></td>
<td></td>
<td>28</td>
</tr>
</tbody>
</table>

Source: authors
Table 2: The 14 SGAs subject to content analysis

<table>
<thead>
<tr>
<th>Location</th>
<th>Date completed</th>
<th>Link to MA</th>
<th>Type of assessment document</th>
<th>Citation†</th>
</tr>
</thead>
<tbody>
<tr>
<td>Coastal British Columbia, Canada</td>
<td>2004</td>
<td>Approved</td>
<td>Final report¹</td>
<td>Coast Information Team (2004)</td>
</tr>
<tr>
<td>Japan (Satoyama–satoumi ecosystems)</td>
<td>2012</td>
<td>Not linked</td>
<td>Summary report*</td>
<td>Duraiappa, Nakamura, Takeuchi, Watanabe and Nishi (2012)</td>
</tr>
<tr>
<td>Northern Range of Trinidad</td>
<td>2005</td>
<td>Approved</td>
<td>Final report</td>
<td>Northern Range Assessment (2005)</td>
</tr>
<tr>
<td>Portugal</td>
<td>2004</td>
<td>Approved</td>
<td>Final report²</td>
<td>Pereira, Domingos, and Vicente (2004). This contributed to the development of the later national level assessment completed after this research was conducted.</td>
</tr>
<tr>
<td>San Pedro de Atacama, Chile</td>
<td>2005</td>
<td>Approved</td>
<td>Executive summary</td>
<td>RIDES (2005)</td>
</tr>
<tr>
<td>Switzerland</td>
<td>2011</td>
<td>Not linked</td>
<td>Assessment framework and methodology</td>
<td>Staub, C., Ott, W. et al. (2011).</td>
</tr>
</tbody>
</table>

* 445 page final report of assessment not analysed as only certain sections are available in English
† As suggested by authors
Source: authors
<table>
<thead>
<tr>
<th>Assessment</th>
<th>Total pages*</th>
<th>No. (%) on CES specifically</th>
<th>No. (%) mentioning tourism</th>
<th>Other measures of tourism content</th>
<th>Main types of tourism considered</th>
</tr>
</thead>
<tbody>
<tr>
<td>Caribbean Sea (CARSEA)</td>
<td>51</td>
<td>15 (29%)</td>
<td>27 (52%)</td>
<td>5 pages of 10-page section assessing the state of ecosystems, focus on tourism. 20 pp of a 29-page annex on scenarios consider tourism.</td>
<td>Mass Niches Coastal Beach Cruise</td>
</tr>
<tr>
<td>Coastal British Columbia, Canada</td>
<td>21</td>
<td>Less than 1 (3%)</td>
<td>0</td>
<td>1 mention of recreation as a cultural feature that can be protected in conservation reserves.</td>
<td>In conservation areas</td>
</tr>
<tr>
<td>Coffee-growing regions of Colombia</td>
<td>30</td>
<td>Less than 1 (3%)</td>
<td>2 (6%)</td>
<td>Very limited discussion of cultural services and ecotourism is only cultural service discussed.</td>
<td>Ecotourism and national parks</td>
</tr>
<tr>
<td>Norway / Glomma</td>
<td>90</td>
<td>20 (22%)</td>
<td>13 (14%)</td>
<td>The products of 6 ecosystem types discussed and recreation is mentioned under all 6.</td>
<td>Alpine Ecotourism Nature-based Outdoor life and recreation (i.e. fishing hunting swimming boating)</td>
</tr>
<tr>
<td>Indian Urban Resource</td>
<td>28</td>
<td>5 (18%)</td>
<td>4 (14%)</td>
<td>Section focussed on ecotourism contains 200 words out of a 3,500 word summary report.</td>
<td>Ecotourism</td>
</tr>
<tr>
<td>Greater Jakarta Bay, Indonesia</td>
<td>23</td>
<td>2 (9%)</td>
<td>16 (69%)</td>
<td>1 page focuses on tourism</td>
<td>Coastal Coral reef Diving</td>
</tr>
<tr>
<td>Location</td>
<td>Total Pages</td>
<td>Pages on Tourism and Recreation</td>
<td>Pages on Other</td>
<td>Notes</td>
<td></td>
</tr>
<tr>
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<td>-------------</td>
<td>---------------------------------</td>
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</tr>
<tr>
<td>Japan (Satoyama-satoumi)</td>
<td>32</td>
<td>5 (16%)</td>
<td>11 (34%)</td>
<td>8 ecosystems services identified and 2 focus on tourism and recreation</td>
<td></td>
</tr>
<tr>
<td>Laguna Lake Basin, Philippines</td>
<td>166</td>
<td>5 (3%)</td>
<td>3 (2%)</td>
<td>Limited discussion of tourism but recreation mentioned on 5 pages</td>
<td></td>
</tr>
<tr>
<td>Northern Range of Trinidad</td>
<td>114</td>
<td>35 (31%)</td>
<td>27 (23%)</td>
<td>Highlights tourism as a key issue</td>
<td></td>
</tr>
<tr>
<td>Portugal</td>
<td>54</td>
<td>5 (9%)</td>
<td>17 (31%)</td>
<td>3 pages on recreation discussed include tourism</td>
<td></td>
</tr>
<tr>
<td>San Pedro de Atacama, Chile</td>
<td>52</td>
<td>7 (13%)</td>
<td>28 (54%)</td>
<td>5 pages devoted to tourism and regulating tourism is 1 of 8 response options discussed</td>
<td></td>
</tr>
<tr>
<td>Southern Africa Millennium Assessment</td>
<td>55</td>
<td>4 (7%)</td>
<td>18 (33%)</td>
<td>Stresses importance of nature-based tourism</td>
<td></td>
</tr>
<tr>
<td>Switzerland</td>
<td>17 (6%)</td>
<td>5 (29%)</td>
<td>Recreation and tourism constitute 5 of the 23 ecosystem services identified</td>
<td>Hunting, Observing wild species, Natural heritage</td>
<td></td>
</tr>
<tr>
<td>-------------</td>
<td>---------</td>
<td>---------</td>
<td>---------------------------------------------------------------------</td>
<td>---------------------------------------------</td>
<td></td>
</tr>
<tr>
<td>United Kingdom.</td>
<td>79 synthesis report (1452 final report of assessment)</td>
<td>5 (6%) synthesis report 65 (4%) final report of assessment</td>
<td>24 (30%) synthesis report 248 (17%) final report of assessment</td>
<td>11 of 124 key findings in the synthesis report discuss tourism</td>
<td></td>
</tr>
<tr>
<td></td>
<td>Mainly habitat-based: Marine, Coastal, Mountain, Woodland, Rural</td>
<td></td>
<td></td>
<td></td>
<td></td>
</tr>
</tbody>
</table>

N* excludes summary opening pages (e.g. table of contents), references, annexes and glossaries

Source: authors